



Technical Improvements to the Greenhouse Gas (GHG) Inventory for California Forests and Other Lands

FINAL REPORT

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EXECUTIVE SUMMARY

In an effort to help reduce changes in climate, the State of California in 2006 enacted the Global Warming Solutions Act. The Act requires the California Air Resources Board (ARB) to set statewide GHG emission limits, to develop regulations to reduce emissions, and to regularly inventory GHG emissions to and removals from the atmosphere. As part of this inventory, the ARB must account for GHG exchanges in forest and rangeland ecosystems. Under a previous agreement with ARB (Agreement #10-778), Battles *et al.* (2014) used Landscape Fire and Resource Management Planning Tools (LANDFIRE) data products to conduct a stock-change assessment and track carbon dynamics on forest, range, and other lands in California. Through this effort, Battles *et al.* (2014) created a GHG Inventory Tool and provided the first spatial estimates of above-ground vegetation carbon stock changes and associated uncertainties for the entire state for natural and working lands sector. However, Battles *et al.* (2014) noted that there were areas that needed additional investigation to further refine California's carbon inventory. Namely,

1. New vegetation classes were introduced in the 2010 LANDFIRE data products and the effect of these changes on carbon stock change estimates needed to be understood.
2. Above ground carbon estimates associated with urban and agricultural landscapes were not included in Battles *et al.* (2014).
3. Investigation was needed to determine the best source of information for estimating above ground dead biomass carbon pools.
4. Undetected growth in the largest forest vegetation classes in LANDFIRE needed to be evaluated and incorporated into the GHG Inventory Tool developed by Battles *et al.* (2014).
5. The extent and distribution of timber harvest and forest management throughout California public and private lands needed to be quantified between 2001 and 2010 – especially for understanding the implications for above ground carbon stock change assessment.
6. Additional refinement was needed to crosswalk LANDFIRE vegetation classes with IPCC landuse class to improve reporting under both typologies.

Consequently, this project was designed to refine above-ground forest, rangeland, and other lands carbon estimates and accounting methods for above-ground biomass originally reported by Battles *et al.* (2014) for the Air Resources Board's periodic California inventory of atmospheric CO₂ removal and greenhouse gas emissions (using a GHG Inventory Tool).

The report is organized around nine steps used to refine Battles *et al.* (2014) GHG Inventory Tool. In step 1 of the project, we reviewed changes in the 2010 LANDFIRE vegetation types and found that changes will not significantly affect the carbon stock change estimates.

In step 2, we used GIS procedures to combine LANDFIRE products which served as a core data layer and lookup table in the updated GHG Inventory tool.

In step 3, we cross-walked corresponding LANDFIRE vegetation types (as combined with vegetation height and cover) with IPCC AFOLU land categories.

In step 4, we conducted an extensive literature review to construct estimates of above ground carbon stocks with associated agriculture and urban landscapes. This information was summarized and ingested into models and geodatabases (in step 8) to refine estimates of carbon stock changes between 2001 and 2010 for California's forests and other lands.

For step 5, we quantified the differences in dead biomass (and carbon) pools when estimated using FIA field data and fuel loading plot data from various study sites, versus when estimated from LANDFIRE's FCCS and FBFM (Scott and Burgan fire behavior fuel model) mapping products. We found that the FCCS fuel behavior model most closely matched field plot and FIA data on dead biomass (and thus carbon) pools.

In step 6, we summarized the distribution and extent of timber management activities that occurred between 1999 and 2012 and estimated carbon stocks in residues and in wood products. We integrated new estimates of carbon in harvested wood products into the updated GHG Inventory Tool.

In step 7, we evaluated FIA data for the period between 2001 and 2010 to account for forest growth that is undetectable in LANDFIRE data products due to how large tree heights are classified. From this assessment we estimated that large tree biomass increased by 6% within the time period of interest. A coefficient was included for the large tree class to account for undetected growth in the carbon stock change assessment.

In step 8 we used summaries and information developed in steps 1 through 7 to update the GHG Inventory Tool. The tool includes database lookup tables, GIS raster layers and geodatabases that are linked together via ArcGIS models. The GHG Inventory Tool was used to complete step 9 – the carbon stock change assessment for 2001, 2008 and 2010.

Using updated information from steps 1 through 7, GHG Inventory Tool in step 8, and using some initial assumptions (that may be changed as ARB further develops and refines the tool for its needs), we preliminarily estimated that between 2001 and 2010, the total above ground carbon stored in the forests, woodlands, shrublands, grasslands, agricultural, developed/urban and other lands of California decreased from 2,696 million metric tons of carbon (MMTC) in 2001 to 2,551 MMTC in 2010, representing a potential overall loss of about -145 MMTC over the time period of interest or a loss of approximately $-16.1 \text{ MMTC yr}^{-1}$. The greatest estimated loss in carbon pools occurred in the form of forest conversion to grassland with wetlands remaining relatively unchanged across 2001 and 2010. These estimates include above ground live biomass associated with forestlands, croplands, grasslands, wetlands, urban/developed (IPCC 'settlements'), and other lands. Stock-changes are reported without attribution by processes such as wildfire or harvest. Stock-changes associated with wildfire and harvest were estimated independently and are provided for informational purposes only. Forestlands represent the largest carbon pool within the study area, storing about 11 times more carbon than other land categories combined. In addition, we preliminarily assessed the carbon stock changes associated with landuse conversions between 2001 and 2010 and found that the largest reduction in net above ground live carbon across wildland, agriculture and urban landscapes was the conversion of the forestland type to the grassland type, and the greatest gain in above ground live carbon was the conversion of the wetland type to the forest type.

INTRODUCTION

In an effort to reduce changes in climate, the State of California in 2006 enacted the Global Warming Solutions Act ([Assembly Bill 32](#)). The Act requires the California Air Resources Board (ARB) to set statewide GHG emission limits, to develop regulations to reduce emissions, and to regularly inventory GHG emissions to and removals from the atmosphere. As part of this inventory, the ARB must account for GHG exchanges in forest and rangeland ecosystems. Vegetation naturally removes GHG's from the atmosphere, reducing the magnitude of climate change. Globally, vegetation and soils removed carbon from the atmosphere at a rate (mean \pm 90% CI) of $2.5 \pm 1.3 \text{ PgC y}^{-1}$ from 2002 to 2011, compared to fossil fuel emissions of $8.3 \pm 0.7 \text{ PgC y}^{-1}$ and deforestation emissions of $0.9 \pm 0.8 \text{ PgC y}^{-1}$ (Table 6.1 in Ciais *et al.* 2013 [i.e., Chapter 6 - IPCC 2013]). Recent estimates for California's forest have varied greatly from a net carbon uptake of 15.7 million MgC y^{-1} (Zheng *et al.* 2011) to net carbon loss of -0.4 million MgC y^{-1} (USFS 2013).

Project Background

Under an agreement with ARB (Agreement #10-778), Battles *et al.* (2014) used U.S. Department of Agriculture Forest Service's and U.S. Department of the Interior's - Landscape Fire and Resource Management Planning Tools (LANDFIRE) data products to conduct a stock-change assessment and track carbon dynamics on forest, range, and other lands in California. Based on their stock-change analysis, which included carbon pools in forests and other lands, except above ground biomass associated with urban and agricultural lands, and soil, Battles *et al.* (2014) reported that between 2001 and 2008, the total above ground carbon stored in the forests and rangelands of California decreased from 2,600 million metric tons of carbon (MMTC = 10^6 MgC) to 2,500 MMTC. Aboveground live carbon decreased ~2% and total carbon (which include carbon associated with dead biomass) decreased ~4%, which represented a statistically significant loss of carbon with an annual rate of approximately -14 MMTC y^{-1} . Battles *et al.* (2014) concluded in general terms that 61% of the loss was due to a reduction in the carbon stored per area (i.e., carbon density), with the remaining 39% due to a reduction in size of the analysis area (i.e., due to wildfire-related transitions of shrublands to grasslands or other land conversions).

Through this effort, Battles *et al.* (2014) created a GHG Inventory Tool and provided the first spatial estimates of above-ground vegetation carbon stock changes and associated uncertainties for the entire state. In doing so, Battles *et al.* (2014) established the beginning of a time series to track above-ground carbon stocks and stock-change in California natural ecosystems. However, Battles *et al.* (2014) noted that there were several areas that needed additional investigation to further refine California's above-ground carbon inventory for forests and other lands, namely:

- Battles *et al.* (2014) relied on land cover metrics provided by LANDFIRE to stratify the state into fine-grained (30m by 30m) spatial units. These metrics, defined by LANDFIRE as Existing Vegetation Type (EVT), Existing Vegetation Cover (EVC) and Existing Vegetation Height (EVH), were subsequently linked by Battles *et al.* (2014) to data on biomass contained in major ecosystem pools (i.e., live vegetation, standing dead vegetation, dead and down wood, litter). The resulting biomass look-up table served as the cornerstone that translated remotely sensed changes in vegetation and land cover to changes in ecosystem carbon (Battles *et al.* 2014).

Based on the 2008 LANDFIRE products, Battles *et al.* (2014) parameterized 1,083 distinct biomass classes (i.e., possible combinations of vegetation type, cover and height classes)

that uniquely assigned carbon densities to every LANDFIRE pixel. That is, every pixel in the analysis area (defined as forests and other natural lands) had a matching biomass class. The assumption was that the land cover classification (i.e., vegetation type, cover and height) for LANDFIRE would remain consistent through time. Indeed there were only minor differences between the 2001 and 2008 LANDFIRE products. However, the 2010 LANDFIRE product made multiple revisions to the vegetation classification system. Namely, of the 200 relevant EVT's in the 2008 biomass lookup table, there were 61 revisions. More than 70% (49 classes) of the changes apply to urban and agricultural lands, parts of the State that lie outside of Battles *et al.* (2014) "Forest and Natural Areas" analysis area. However, there were 10 new categories that more finely divided types classified as "recently disturbed developed uplands" in 2008 to developed and undeveloped "ruderal" vegetation types. The relevance to the biomass look-up table was that there were now 10 new vegetation types in the analysis area – all the 2010 types defined as "undeveloped ruderal." However, there are no estimates for "new" vegetation classes in the Battle *et al.* (2014) biomass classes look-up table because they were considered part of the urban footprint by the 2008 classification. The remaining 12 revisions all involved tree-dominated types and collectively include 15% of the forest lands. The most significant change in terms of carbon storage was the division of the 2001/2008 California Lower Montane Blue Oak-Foothill Pine Woodland and Savanna into three separate EVT's in the 2010 LANDFIRE data product. This EVT is one of the most common in California (12.7% of forest land) and contains on average approximately 40 MgC/ha in the live vegetation (Battles *et al.* 2014). In a similar fashion, the 2001/2008 Mediterranean California Lower Montane Black Oak-Conifer Forest and Woodland (1.9% of forest land; on average 94 MgC/ha in the live vegetation) was revised into three separate EVT's in 2010. The remaining six revisions involved rare types (<0.3% of forest area). Consequently, there was a need to investigate the implications of changes to EVT for the state GHG Inventory Tool.

- The Battles *et al.* (2014) analysis did not include estimates of carbon stocks and stock change associated within agricultural and developed (i.e., urban) landscapes. A review of literature and other sources of information was needed to update the biomass classes lookup table (noted above) in order to improve estimates of carbon-stock changes associated of these land types.
- The reliance on LANDFIRE vegetation height data layer (Existing Vegetation Height or EVH) limited the resolution at which tree growth (and associated carbon sequestration) could be detected, particularly for mature forests where the range in tree height categories is greater for mature forest than for younger forest. As a result, Battles *et al.* (2104) generally concluded that the method used likely underestimated live tree carbon densities for the most carbon dense forest types in California and the means to account for undetected growth was needed to improve estimates of annual carbon pools and associated stock change.
- Until 2012 the US Forest Service Forest Inventory and Analysis program (FIA) used a model to generate estimates of specific carbon pools, including dead carbon, in compliance with IPCC recommendations. These estimates were incorporated into Battles *et al.* (2014) GHG Inventory Tool. However, information associated with the fuelbeds of the Fuel Characteristic Classification System (FCCS; Ottmar *et al.* 2007, Prichard *et al.* 2013) and with Scott and Burgan's (2005) fire behavior fire models provide other tools which can provide estimates of carbon pools, including dead biomass carbon associated

with litter, duff and coarse woody debris. The primary purpose of FCCS is to quantify dead and live fuels within several strata into meaningful categories for predicting fire behavior and emissions. Hundreds of FCCS fuelbeds, which quantify fuels in the different strata (overstory, understory, litter, etc.) have been developed, and subsequently mapped across the United States as a component of the LANDFIRE program. As an enhancement to the Battles *et al.* (2014) GHG Inventory Tool, FCCS estimates of dead carbon pools were included for the forest/range/other natural lands – where an analyst using the tool can include either FIA or FCCS information. However the choice leads to differences in the statewide carbon flux estimates. One major difference is that the IPCC definition (and FIA) of litter is more inclusive than the FCCS definition of “litter and duff.” Thus, the magnitude of the flux tends to be greater with FIA estimates.

Scott and Burgan (2005) developed fire behavior fuel models primarily for modeling fire behavior but they could also serve as an alternative to FIA- and FCCS-based carbon estimation. However, the value of Scott and Burgan (2005) for carbon estimation was unknown and not investigated in Battles *et al.* (2014). Consequently, an identified need for refinement to the GHG Inventory Tool was to quantify and understand the differences in dead carbon pools when estimated using FIA field data and non-FIA field data versus when estimated by FCCS and Scott and Burgan (2005) fuelbeds. Results of such an analysis could be used to identify the more appropriate means to assess dead wood carbon pools within the context of the Battles *et al.* (2014) GHG Inventory Tool.

- Although Battles *et al.* (2014) included emission estimates from timber management activities (i.e., silviculture applications and practices) and logging residuals (see Appendix 3 in Battles *et al.* 2014), its accounting of carbon stored in harvested wood products followed simplified life cycle assessment scenarios (e.g., DOE 2007 guidelines) and did not include harvests on public lands. Quantifying the carbon stock changes associated with management of forests and other vegetation types is complicated by variations in the intensity of activities such as timber harvesting (including both commercial and non-commercial operations) according to ownership type and the fate of wood products and residuals. Additionally, vegetation management and harvest activities lead to carbon stock changes that are difficult to quantify using remotely-sensed data since such activities are periodic in nature and often do not coincide with remote sensing production dates. Carbon stocks at treatment sites can recover at varying rates in between data acquisition years, and variation in accounting for the fate of carbon in harvested wood products and residuals can confound the assessment of carbon stock changes. Assessing stock changes from vegetation management and harvest activities requires calibration between site-level removal (harvest) data and remotely-sensed data that is adjusted for land ownership type, the temporal lag of monitoring data, and adjusting for the fate of wood products and residuals. Consequently, more investigation was needed to account for carbon associated with harvested wood products and estimates of carbon stock changes associated with timber harvest and management.
- Battles *et al.* (2014) used LANDFIRE data products to classify land cover types and associated carbon pools within pixels across the state of California. However, ARB needs to report information in a variety of formats, including standard IPCC categories for the Agriculture, Forestry and Other Land Use (AFOLU) sector, as well as custom formats and categories defined by ARB. The IPCC generally defines six broad categories of land for reporting on AFOLU, these are: Forestland, Cropland, Grassland, Wetland, Settlements,

and Other Uses. A variety of issues can come up when assigning land cover classes to these broad categories. The issues typically arise at the point where thresholds of vegetative cover (such as projected canopy cover) must be defined and land cover classes have attributes of two IPCC categories such as Forestland and Grassland. This is particularly important in the managed forestland context where stock-change occurs for short time periods but the functional definition of forest is more relevant than assigning a land use change value to the area. Other such classification decisions need to be made at the boundary of Wetlands-Grassland and Cropland-Grassland (and to some degree Forestland-Cropland where woody nut-tree crops are dominant).

Because of these outstanding investigation needs, ARB set out to further refine the state carbon inventory program through this project.

Physical and Operational Boundaries (Scope)

One of the first steps in preparing a GHG inventory is to define physical and operational boundaries (i.e., scope) of the inventory. A definition of physical boundary typically includes the spatial extent of the inventory, for example for the Battle *et al.* (2014) GHG Inventory Tool the boundaries were the state of California. The operational boundaries define which direct and indirect emissions (losses) and removals (gains/sinks/pools) that are included in a GHG inventory. For operational boundaries, Battles *et al.* (2014) evaluated above-ground carbon pools and stock change associated with tree, shrub and herbaceous vegetation dominated landscapes, including forests, shrublands, grasslands, wetlands and desert habitats. The inventory included estimates for the carbon stored in both live and dead vegetation pools. The Battles *et al.* (2014) study did not include an evaluation of below ground carbon pools, or soil and above-ground biomass carbon pools associated with developed (urban) or agricultural lands. The scope of the GHG Inventory Tool as updated through this project includes:

Carbon Stocks and Stock Gains

- Above ground live biomass
 - Biomass associated with forest, rangelands, wetlands, desert and other natural lands (undeveloped or not cultivated)
 - Forest vegetation
 - Shrub vegetation
 - Herbaceous vegetation
 - Above ground live biomass associated with developed/urban lands and settlements
 - Urban “forests” and trees
 - Urban shrub vegetation
 - Urban herbaceous vegetation
 - Above ground live biomass associated agriculture and cultivated lands
 - Woody/Orchard/Vineyard Crops (e.g., almond, orange, grapes, peaches, etc.)
 - Annual shrub crops
 - Annual herbaceous crops (wheat, broccoli, lettuce, etc.)
- Above ground dead biomass
 - Forest, rangelands, wetlands and other natural lands (undeveloped or not cultivated)
 - Standing dead (snags)

- Course woody debris
- Litter
- In-use wood products (e.g., building materials, furniture, etc.)
- Above ground dead biomass associated with Developed/Urban Lands/Settlements was NOT included
- Agriculture and Cultivated Lands
 - Post-harvest residues (for certain crop types only)

Carbon Losses

- Natural processes – decomposition of biomass, and biomass respiration
- Wildfire (live and dead biomass combustion)
- Timber Harvest and Management
 - Harvest residue emissions on-site
 - Prescribed fire (biomass combustion)
 - Timber harvest and wood products processing emissions
 - Post-use wood products

Only above ground carbon gains and losses associated with biomass are accounted for with the GHG Inventory Tool, soil carbon is not included.

Objectives

The goal of this project was to refine carbon estimates and accounting methods originally reported by Battles *et al.* (2014). To achieve this goal, we conducted applied research with the following objectives:

1. Evaluate and update Battles *et al.* (2014) biomass classes look-up table and geoprocessing procedures to account for vegetation categories contained in the 2010 LANDFIRE Existing Vegetation Type (EVT) data product (LF_1.2.0).
2. Quantify differences generated by USDA Forest Service, Forest Inventory and Analysis (FIA) based estimates of dead carbon pools and non-FIA plot data with Fuel Characteristics Classification System (FCCS)- and Scott and Burgan (2005)- based estimates for key forest and woodland types, including an analysis of the methods underpinning the estimates. Use results to identify options for including dead wood carbon pool estimates into carbon stock change assessment.
3. Conduct a comprehensive review of available information regarding ecosystem carbon stocks for agricultural and developed (urban) landscapes. Compile the information and use it to construct best- available estimates of carbon stock-change associated with conversion of natural landscapes to agricultural or other developed land uses in California.
4. Combine geospatial information on vegetation management and harvest activities from federal agencies with the (UC Berkeley) statewide ecosystems stock-change assessment, to make probability-based assignments of stock-change associated with activities on federal lands.
5. Review assignments of Battles *et al.* (2014) California vegetation cover types to IPCC AFOLU categories based on national practices. Make recommendations on assignment options and quantify the impact of revisions to the statewide ecosystems carbon stock-change assessment.

Report Organization

The body of this report describes the steps we used to update the Battles *et al.* (2014) GHG inventory tool and account for changes to the 2010 LANDFIRE data products. The report is organized around the following steps (see also Figure 1):

1. Review and evaluate the effect of new vegetation categories represented in the 2010 LANDFIRE EVT data layer on biomass and carbon estimates. Identify how to improve consistency of vegetation categories (types) across evaluation years (2001, 2008 and 2010) to facilitate stock change evaluation.
2. Combine 2010 LANDFIRE EVT layer with 2010 EVC and EVH layers to create a new accounting layer and attribute table. The merging of these datasets allowed for the allocation of biomass and carbon estimates for each combination of vegetation type, height and cover class and was used to show how these combinations of classes are distributed across California's landscape.
3. Crosswalk IPCC land category typologies (i.e., forestland, cropland, grassland, wetlands, settlements and other lands) with geospatial accounting layer categories (as derived from LANDFIRE). This step was needed to translate and communicate the classification scheme used for the GHG Inventory Tool with IPCC land categories – allowing estimates of biomass and carbon pools and stock change to be reported under different reporting schemes, including typologies that ARB may choose to use in the future.
4. Conduct literature and data review, and summary of biomass and carbon associated agriculture and urban landscapes. This step was needed because these two landuse types were not evaluated in Battles *et al.* (2014) carbon stock change estimates and were needed to gain greater understanding of their role in accounting for California's above ground carbon pool. Information from this review was used to update the biomass classes lookup table and to include biomass and carbon estimates into the GHG Inventory Tool.
5. Evaluate dead carbon pools associated with fuelbeds and identify best fuel bed/model or data option for potential use in updated biomass classes lookup table and GHG Inventory Tool.
6. Evaluate distribution and extent of timber management activities that occurred between 1999 and 2012 and integrate into the updated GHG Inventory Tool for considerations on the persistence of carbon in harvested wood products. This step was needed to provide a methodology for allocating and quantifying timber harvest related carbon retention (in wood products) and losses (emission) on the landscape across California from 2001 to 2010.
7. Evaluate available data to determine best option to account for undetected biomass growth in LANDFIRE data products. Incorporate growth estimates into updated GHG Inventory Tool.
8. Assign biomass and carbon estimates for new LANDFIRE vegetation categories (and associated IPCC land use categories, from steps 2 and 3), agriculture and urban (from step 4), dead wood (from step 5), timber management (from step 6) and undetected growth (from step 7) into updated 2010 biomass classes lookup table and accounting layer.
9. Conduct stock change analysis for 2001 and 2010 using updated biomass classes lookup table and GHG Inventory Tool.

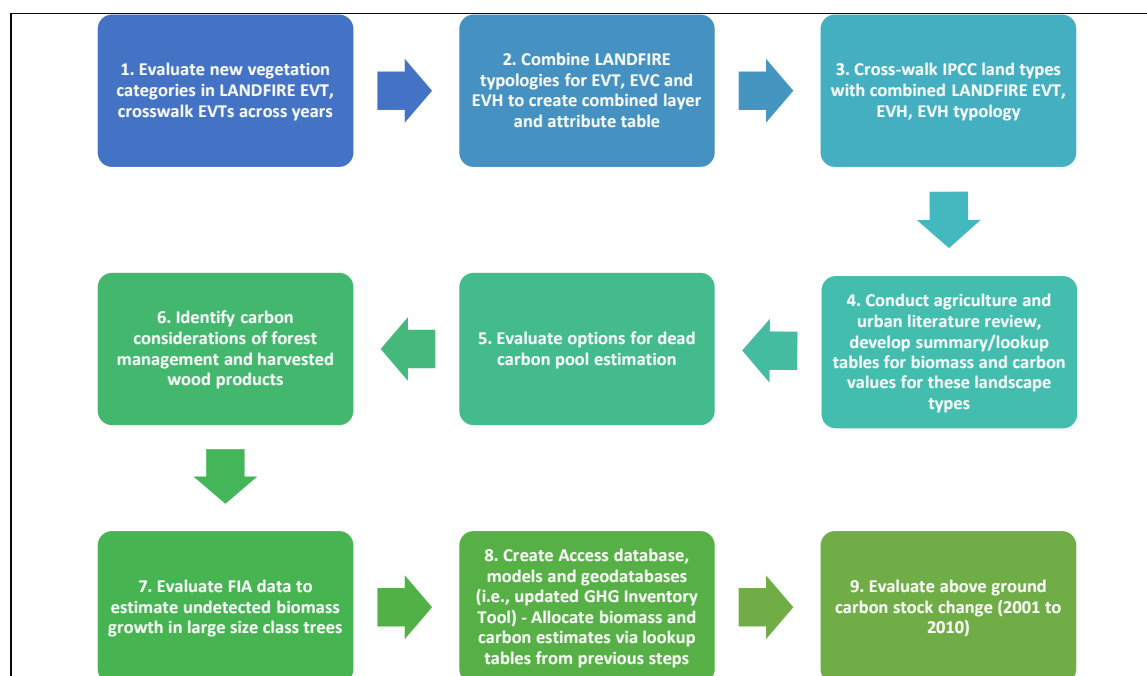


Figure 1. Steps used to update Battle *et al.* (2014) GHG Inventory Tool.

METHODS AND RESULTS

Step 1: Review of 2010 LANDFIRE Vegetation Type Category Changes

A key goal of the LANDFIRE program is to provide a consistent national vegetation map that is sufficiently resolved to inform decisions about resource management and policy. In an effort to remain consistent with National Vegetation Classification Standards (NVCS 2015), LANDFIRE follows a hierarchical system with the most general category (Order) defined by the form of the dominant vegetation: tree, shrub, herb, no dominant lifeform, and no vegetation (Table 1). Subsequent levels include ‘class’ where the dominant vegetation is modified by its gross structure. For example, classes within the order of ‘tree’ include ‘closed-canopy’, ‘open canopy’, and ‘sparse-tree canopy’. The ‘subclass’ divides canopy structure by leaf form. For example, the class of closed-canopy tree is separated into ‘evergreen’, ‘deciduous’, or ‘mixed’. The most finely resolved vegetation category is the ‘existing vegetation type’ (EVT). This LANDFIRE category is equivalent to the sub-regional NVCS definition of a ‘group’ (Table 1), defined as: “A *vegetation classification unit of intermediate rank (6th level) defined by combinations of relatively narrow sets of diagnostic plant species (including dominants and co-dominants), broadly similar composition, and diagnostic growth forms that reflect biogeographic differences in mesoclimate, geology, substrates, hydrology, and disturbance regimes* (FGDC 2008).”

Table 1. Vegetation classification levels, classification criteria and examples of the levels of the National Vegetation Classification Standard hierarchy for natural vegetation.¹

Vegetation Classification Level	Vegetation Classification Criteria	Ecological Context	Scientific Name	Common Name
Upper Levels	Predominantly physiognomy			
1. Formation Class	Broad combinations of general dominant growth forms.	Basic temperature (energy budget), moisture, and substrate/aquatic conditions.	Mesomorphic Tree Vegetation	Forest and Woodland
2. Formation Subclass	Combinations of general dominant and diagnostic growth forms.	Global macroclimatic factors driven primarily by latitude and continental position, or overriding substrate/aquatic conditions.	Temperate Tree Vegetation	Temperate Forest
3. Formation	Combinations of dominant and diagnostic growth forms.	Global macroclimatic factors as modified by altitude, seasonality of precipitation, substrates, and hydrologic conditions.	Cool Temperate Tree Vegetation	Cool Temperate Forest
Middle Levels	Physiognomy, biogeography, and floristics			
4. Division	Combinations of dominant and diagnostic growth forms and a broad set of diagnostic plant species that reflect biogeographic differences.	Continental differences in mesoclimate, geology, substrates, hydrology, and disturbance regimes.	Pseudotsuga - Tsuga - Picea - Pinus Forest Division	Western North America Cool Temperate Forest
5. Macrogroup	Combinations of moderate sets of diagnostic plant species and diagnostic growth forms that reflect biogeographic differences.	Sub-continental to regional differences in mesoclimate, geology, substrates, hydrology, and disturbance regimes.	Pseudotsuga menziesii - Quercus garryana - Pinus ponderosa - Arbutus menziesii Macrogroup	Northern Vancouverian Montane and Foothill Forest
6. Group	Combinations of relatively narrow sets of diagnostic plant species, including dominants and co-dominants, broadly similar composition, and diagnostic growth forms.	Regional mesoclimate, geology, substrates, hydrology and disturbance regimes.	Pinus ponderosa - Quercus garryana - Pseudotsuga menziesii Group	East Cascades Oak-Ponderosa Pine Forest and Woodland
Lower Levels	Predominantly floristics			
7. Alliance	Diagnostic species, including some from the dominant growth form or layer, and moderately similar composition.	Regional to subregional climate, substrates, hydrology, moisture/ nutrient factors, and disturbance regimes.	Pinus ponderosa - Quercus garryana Woodland Alliance	Ponderosa Pine - Oregon White Oak Woodland Alliance
8. Association	Diagnostic species, usually from multiple growth forms or layers, and more narrowly similar composition.	Topo-edaphic climate, substrates, hydrology, and disturbance regimes	Pinus ponderosa - Quercus garryana / Balsamorhiza sagittata Woodland	Ponderosa Pine - Oregon White Oak / Arrowleaf Balsamroot Woodland

¹ Source: <http://usnvc.org/data-standard/natural-vegetation-classification/>

In this carbon stock assessment, we took advantage of the mesoscale resolution (As defined by LANDFIRE) of the LANDFIRE EVT's to assign biomass values (Battles *et al.* 2014). Since the EVT is determined by the dominant vegetation, it was no surprise that EVT proved to be the best single predictor of aboveground live biomass for forests and other working lands in California. Thus our system relies on a consistent determination of EVT as LANDFIRE updates land cover and land use change through time. However dynamic mapping of vegetation for the entire United States requires the means to process several hundred thousand vegetation plots and apply labels matching the EVT definitions. By their own admission, there was limited time to evaluate the performance of the mapping algorithms (referred to as "auto-keys"). Moreover, the baseline LANDFIRE classification system itself has changed over time in order to match revisions to the NVCS. As a consequence, the EVT designations are not consistent as LANDFIRE is updated over time.

This inconsistency requires a cross-walk between EVT's for every mapped iteration of LANDFIRE in order to assess stock changes in carbon. Indeed we did this for the 2001 to 2008 analysis in Battles *et al.* (2014) and again for the 2001 to 2010 analysis in Gonzalez *et al.* (2015). These cross-walks were based on the matching descriptions of the EVT using the dominant species, the vegetation structure, and edaphic qualifiers. Elsewhere in this report, we provide a comprehensive crosswalk to biomass look-up tables for all EVT classes (including classes associated with agriculture and urban landscapes) for every LANDFIRE version (2001, 2008, and 2010). Here we explored how revisions in the LANDFIRE vegetation mapping may impact carbon stock assessment.

Carbon Implications of EVT Assignment. The majority of changes in LANDFIRE EVT's are the result of efforts to more finely resolve vegetation classes. Thus over time, there are more EVT classes (Table 2). The reason for these fall into two categories: 1) For EVT's with shared dominance between deciduous and evergreen trees, the 2010 revision separated the EVT into two classes based on tree composition and 2) for EVT's that included more than one vegetation structure (e.g., forest and woodland), the 2010 class was divided into two based on vegetation structure. Other revisions were more of a book-keeping nature. For example in 2010, some EVT's that included the common name of the dominant species in the name were changed to the scientific name (e.g., the Douglas-fir-Oregon White Oak Woodland became the *Pseudotsuga menziesii-Quercus garryana* Woodland Alliance). While it is a chore to account for such name changes, they will not affect the carbon estimates. In contrast, the division of EVT's by species composition or vegetation structure might provide more refined categories for biomass assignments, particularly when an abundant or carbon dense EVT is split.

Table 2. Number of existing vegetation types (EVT's) by major landuse type in California from three LANDFIRE data iterations (2001, 2008 and 2011).

LANDFIRE Year/Version	Irrigated Agriculture	Urban	Forests and Working Lands	Total
2001	9	10	138	158
2008	16	10	141	168
2010	30	19	154	204

A good test case for California is the “California Lower Montane Blue Oak-Foothill Pine Woodland and Savanna” (i.e., blue oak woodlands). It is an example of an EVT that was split into three separate EVT’s in 2010 based on compositional differences (oak dominance versus pine dominance) and structure (forest/woodland versus savanna). It is the second most common vegetation type (by area) in the state covering 16,740 km² and accounting for 8% of the above-ground carbon stock (based on 2008 LANDFIRE - urban and irrigated agricultural lands excluded, Battles *et al.* 2014). Based on FIA plot data for this EVT (277 plots), carbon density varies by an order of magnitude. Below three specific objectives are outlined to quantify the carbon implications of the revisions to the blue oak woodland EVT.

Objective 1 - The EVT Classification Process. In 2010, LANDFIRE divided the blue oak woodland into three separate EVT’s: California Lower Montane Blue Oak Forest and Woodland, California Lower Montane Blue Oak-Foothill Pine Forest and Woodland, California Lower Montane Foothill Pine Woodland and Savanna. This revision relies on the LANDFIRE mapping algorithm (auto key) to parse the previous EVT into a more tree-centric, oak dominated class from a more open, savanna class dominated by pines while also retaining a mixed species designations for the sites in the middle of this gradient. LANDFIRE 2010 adopted these divisions even though the NatureServe analysis on mapping accuracy specifically notes the difficulty of distinguishing floristically similar ecological systems and the gains in accuracy associated with slightly coarser vegetation classes (NatureServe 2012). Indeed, the auto key results for the coarser 2008 EVT matched expert opinion 84% of the time (42 correct out of 50, NatureServe 2012). There was no accuracy assessment conducted for the revised 2010 LANDFIRE EVT’s.

Objective 2 – Quantify how well the species compositional differences in the 277 FIA plots classified in 2008 as California Lower Montane Blue Oak-Foothill Pine Woodland and Savanna predict differences in aboveground live biomass. We analyzed the gradients in species composition for the 277 FIA plots classified by LANDFIRE 2008 in the coarse blue oak woodland EVT (i.e., California Lower Montane Blue Oak-Foothill Pine Woodland and Savanna). We used Detrended Correspondence Analysis (DCA) to quantify gradients in species composition and regression tree analysis to determine how well these gradients predicted aboveground live biomass (ALB; McCune *et al.* 2002). As expected we detected a significant compositional gradient between plots with more oaks and plots with more conifers. However, neither blue oak nor foothill pine dominance was a robust predictor of ALB (Table 3). The best predictor of ALB was the abundance of Douglas-fir, a relatively minor determinant of the compositional gradient.

Table 3. Comparison of relative variable importance (rVIP) for determination of floristic classification versus determination of carbon density for 277 blue oak woodland FIA plots.

Species Dominance	Floristics (rVIP%)	Carbon density (rVIP%)
Blue oak	47	7
Live oaks	30	10
Black oak	6	8
Foothill pine	5	1
Other pines	5	3
Douglas-fir	4	57
California juniper	2	1
Coast redwood*	1	11

*Note: plots with coast redwood represent misclassified plots by LANDFIRE (2008).

Objective 3 - Quantify how well vegetation structure differences in the 277 FIA plots classified in 2008 as California Lower Montane Blue Oak-Foothill Pine Woodland and Savanna predict differences in aboveground live biomass. We used the primary division in the regression tree results to divide the plots into low biomass ($n=130$) and high biomass groups ($n=147$). Given that the definition of savanna compared to woodland (Allen-Diaz *et al.* 1999) implies less tree cover, we assigned the low biomass plots as savanna and the high biomass plots as forest/woodland. On average, the mean ALB for forest/woodland group was 41.1 MgC/ha (interquartile range: 17.7–56.4 MgC/ha) and for the savanna plots, 17.8 MgC/ha (interquartile range: 8.2 – 24.3 MgC/ha). ALB regression equations were fit for each group as functions of cover and height class following the same model selection criteria used in Battles *et al.* (2014). The results were two submodels: one for the forest/woodland plots and one for the savanna plots.

The resulting transfer functions clearly captures the disparities with the forest/woodland predicting higher ALB for each cover/height combination (Figure 2). To quantify the impact compared to the inclusive model (all plots, no separation by structure), we subtracted the inclusive model estimates (all 277 plots) from the submodel estimates (Figure 3). On average the forest/woodland submodel resulted in 16% higher estimates of ALB and the savanna submodel in 11% lower estimates. Interestingly, for all cover/height class combinations, the forest/woodland submodel produced higher ALB estimates. For the savanna submodel, differences ranged from positive for lower cover classes to negative for the higher cover classes. Despite the differences in the two submodels, the increase in precision is modest relative to the overall variability in ALB in the blue oak woodlands. While the models were robust and captured the trends in ALB with cover and height ($R^2 > 0.85$), the relative error of the estimate was $> 40\%$.

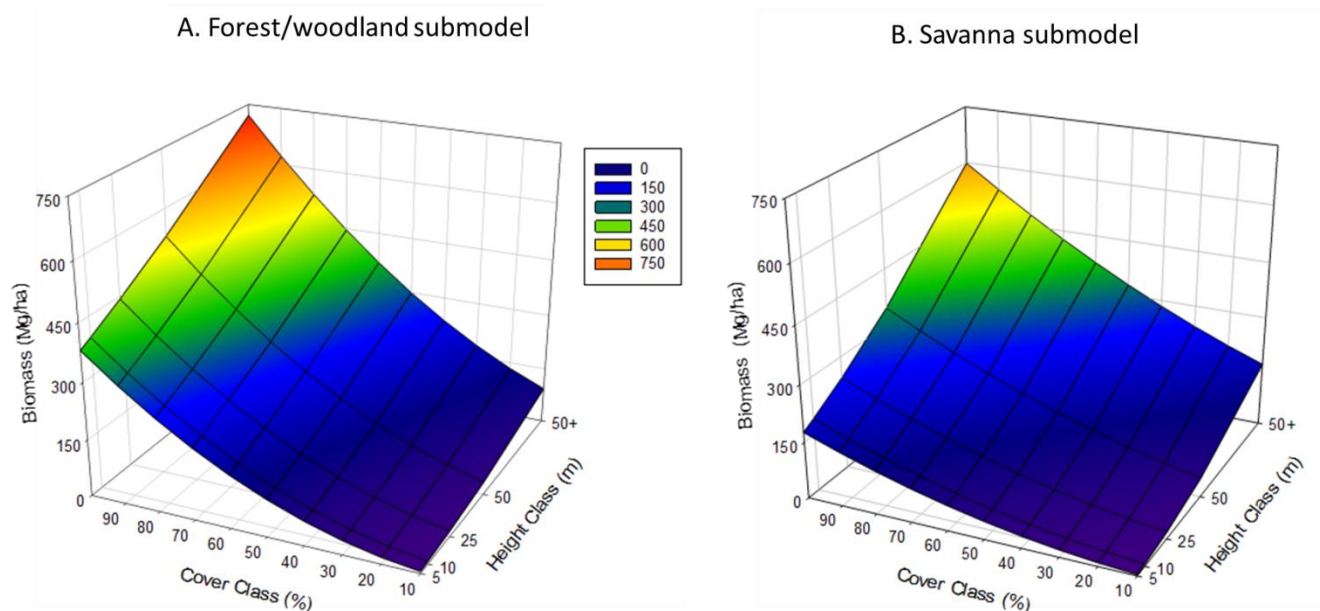


Figure 2. Predicted variation in ALB as a function of cover and height class. A) Results for the forest/woodland submodel; B) Results for the savanna submodel.

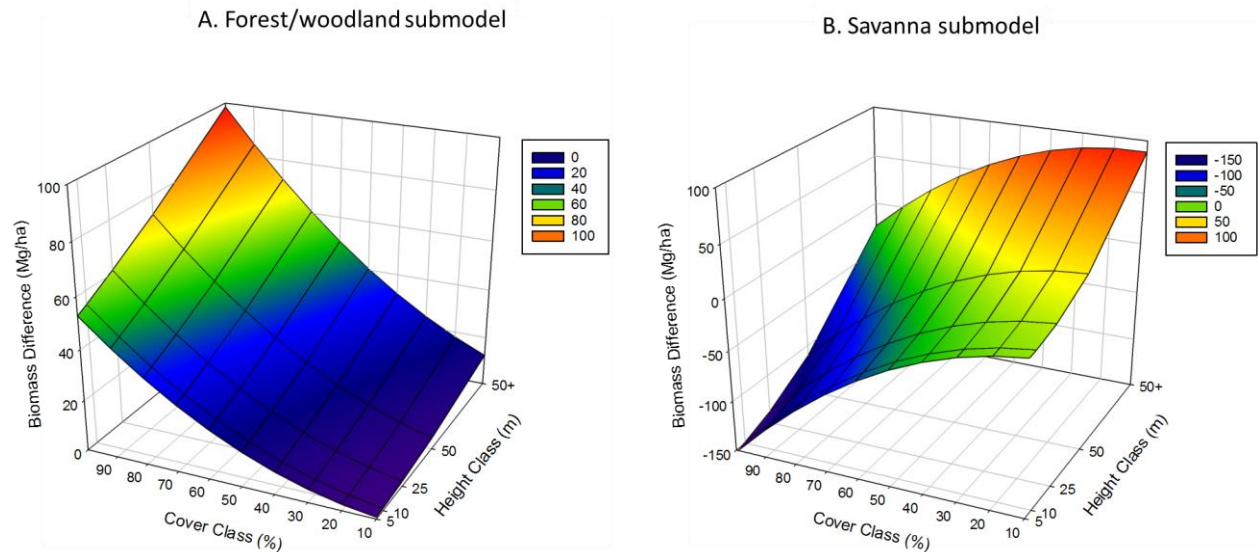


Figure 3. Differences in predicted ALB for each submodel compared to inclusive model. A) Results for the forest/woodland submodel; B) Results for the savanna submodel.

Step 2: Integration of LANDFIRE EVT, EVC and EVH Data Layers

The vegetation classification system used by LANDFIRE, its fundamental logic and its evolution over the three LANDFIRE revisions directly informed the integration of LANDFIRE EVT, EVC, and EVH data layers. As noted above, EVT is the primary layer for predicting biomass storage and thus consistent EVT assignments by the LANDFIRE program through time is key. In most cases, we relied on the description of the dominant vegetation to match shifts in the EVT definition. For example, the three “new in 2010” EVT’s that subdivided the California Lower Montane Blue Oak-Foothill Pine Woodland and Savanna were all assigned to the same biomass class because the dominant species (blue oak and foothill pine) are contained within the 2008 designation. For newly defined classes without a dominant vegetation assignment (e.g., California Central Valley Riparian Forest and Woodland), we assigned the biomass class associated with the likely subclass designation of the dominant of vegetation. For the Central Valley, the riparian forests are dominated by cottonwoods and willows that form a relatively open canopy structure. Hence we assigned the biomass associated with the vegetation subclass, “deciduous open-tree canopy.” Once the EVT was assigned, the estimate biomass storage was further parsed by the attendant EVC and EVH designations. In summary, raster LANDFIRE data layers (EVT, EVH and EVC) from 2010 were combined to create a geodatabase - “ARB_LFc.gdb.” The resulting data layer’s (i.e., “ARB_LFc.gdb”) attribute table served as a foundation for the updated 2010 biomass classes lookup table for which estimates of biomass density (Mg/ha) were applied to each biomass class (e.g., Mediterranean California Mixed Evergreen Forest, Forest Height 0 to 5m, Tree cover ≥ 20 to $<30\%$). We also added the new biomass classes to account for urban and agricultural lands (described in steps below). Appendix 1 contains a list of files developed for this project (for all tasks), including the geodatabase and associated attribute table that houses the combined LANDFIRE data (i.e., EVT, EVC, and EVH).

Step 3: Crosswalk IPCC Land Categories with 2010 Carbon Accounting Layer Categories.

This step was relatively straightforward. Each LANDFIRE EVT for every product year (2001, 2008, and 2010) was assigned an Intergovernmental Panel on Climate Change (IPCC) - Agriculture,

Forestry and Other Land Use (AFOLU) category based on the description of the existing vegetation. Corresponding IPCC AFOLU categories and LANDFIRE vegetation type categories (as combined in step 2 with EVC and EVH) were aligned based on their respective definitions and organized into a crosswalk table ("BATTLES_Biomass-LUT_01-08-10_20151029") in the "ARB_C_LUT_v2.7.accdb" database to facilitate queries for either land typology.

Step 4: Literature and Data Review and Summary of Biomass and Carbon Associated Agriculture and Urban Landscapes.

Agriculture and Urban Vegetation Types

We conducted an extensive review of best available science to construct estimates of above ground carbon stocks with associated agriculture and urban landscapes. Literature and data sources consulted included: Google Scholar, Web of Science, UC Agricultural Extension, and local agricultural cooperatives. A Microsoft Access database titled "*ARB literature review database*" was used to organize summarized information including a complete the list of information sources (see Appendix 2) Biomass and carbon estimates extracted from reviewed information are organized into an updated biomass classes lookup table and categorized into one of the corresponding LANDFIRE Existing Vegetation Types associated with agricultural or other developed lands, as well as into corresponding IPCC AFOLU categories (Table 4).

LANDFIRE EVT's for agriculture and urban vegetation categories occur as both "Western Cool Temperate" and "Western Warm Temperate" within California. In most cases, it was not possible to distinguish biomass or carbon stock values in warm vs. cool types, so except where noted the same values are used for each. For some LANDFIRE EVT's, a single value was obtained or calculated to use in the statewide lookup table. Where multiple crop types comprised a LANDFIRE EVT, the values were weighted based on acreage summaries available through the 'CropScape' database (Boryan *et al.* 2011).

Biomass and Carbon Stock Value Calculation Methods

We used five different methods to summarize and estimate biomass and calculate carbon (C) content as the data were presented in different ways in the literature and in relevant databases. Different methods were required because we did not find total tree or plant biomass estimates nor were essential carbon equation parameters available on every crop grown in the state. Table 5 summarizes the general approaches used to quantifying total aboveground biomass for each LANDFIRE EVT. A detailed description of each method as it applies to the LANDFIRE EVT follows. Data sources varied from published literature to online databases (see Appendix 2 for list of information sources).

Vineyard and Orchard Existing Vegetation Types

Vineyard and orchard EVT's included almonds, avocados, oranges, and grapes (Table 6). Estimates of the carbon content of almond (DeJong 2013), orange (Morgan *et al.* 2006), and avocado (Rosecrance and Lovatt 2003), orchards and grape vineyards (Carlisle *et al.* 2010) were made using published data on whole tree or vine biomass estimates, and multiplied by typical planting densities of given species (trees/hectare), and the standard carbon coefficient of 0.47gC/g biomass (McGroddy *et al.* 2004). These were the only crops where this type of data were found and estimated in this way (see also Table 7).

Table 4. Source of biomass and carbon values assigned to different LANDFIRE existing vegetation types (EVT) and IPCC AFOLU categories. Values were either sources from existing literature or databases, calculated using accepted methods or drawn from IPCC Tier 1 default values.

LANDFIRE Existing Vegetation Type (includes both “warm” and “cool” types)	IPCC AFOLU Category	Value Sourced from Literature or Database	Value Calculated	IPCC Tier 1 Default Value
Western Temperate Aquaculture	Cropland			X
Western Temperate Bush Fruit and Berries	Cropland			X
Western Temperate Close Grown Crop	Cropland		X	
Western Temperate Developed Ruderal Deciduous Forest	Settlement			X
Western Temperate Developed Ruderal Evergreen Forest	Settlement			X
Western Temperate Developed Ruderal Grassland	Settlement	X		
Western Temperate Developed Ruderal Mixed Forest	Settlement			X
Western Temperate Developed Ruderal Shrubland	Settlement			X
Western Temperate Fallow/Idle Cropland	Cropland			X
Western Temperate Orchard	Cropland	X		
Western Temperate Pasture and Hayland	Grassland		X	
Western Temperate Row Crop	Cropland		X	
Western Temperate Row Crop-Close Grown Crop	Cropland		X	
Western Temperate Undeveloped Ruderal Deciduous Forest	Forestland			X
Western Temperate Vineyard	Cropland	X		
Western Temperate Wheat	Cropland		X	
Developed Forest	Settlement	X		

Table 5. Summary of evaluation approach used to calculate aboveground biomass and carbon estimates.

Evaluation Category	Method Used to Estimate Carbon Content:
Whole tree/plant above ground crop biomass	Multiplied by typical planting densities of specific crops and the standard carbon coefficient of 0.47gC/gram biomass.
Total yield biomass data per crop	Used Equation 1 (see <i>below</i>) for total biomass estimate and multiplied by the standard carbon coefficient of 0.47gC/g biomass.
Crop residue and total yield biomass estimates	Used Equation 2 (see <i>below</i>) for total biomass estimate and multiplied by the standard carbon coefficient of 0.47gC/g biomass.
Urban Biomass	US Forest Service Forest Inventory and Analysis (FIA) and iTree data summary. Summarized existing urban forest carbon stock data by county (mean MgC/ha)
Value reported directly in literature	Used value without modification or average values if multiple values were reported for a given type.

Table 6. Estimated carbon content of 2014 peak yields of common agricultural commodities of California (National Agricultural Statistics Service 2015).

Commodity	Hectares Harvested (2014)	Estimated MgC/ha
Apples	6,070.5	7.52
Apricots	3,844.7	6.17
Blueberries	1,942.6	5.41
Grapefruit	4,047.0	14.12
Lemons	18,616.2	16.54
Nectarines	8,498.7	9.72
Oranges-Navel	53,825.1	11.58
Oranges-Valencia	14,569.2	12.10
Peaches-Clingstone	8,094.0	17.55
Peaches-Freestone	9,712.8	12.69
Pears-Excl Bartlett	1,052.2	14.28
Pears-Bartlett	3,440.0	19.14
Plums	7,284.6	6.64
Raspberries-Black	283.3	4.37
Raspberries-Red	2,752.0	9.59
Strawberries	16,795.1	35.04
Tangerines	18,211.5	12.73
Almonds	352,089.0	1.01
Avocados	21,772.9	3.23
Olives	14,973.9	2.69
Pistachios	89,438.7	1.10
Walnuts	117,363.0	2.08
Artichokes	2,954.3	6.85
Asparagus	4,451.7	1.63
Broccoli	49,373.4	8.69
Cabbage	6,637.1	22.13
Carrots	26,507.9	16.86
Cauliflower	13,719.3	9.75
Celery	11,007.8	33.72
Cucumbers	1,537.9	9.48
Melons-Cantaloupe	14,569.2	13.70
Melons-Honeydew	4,249.4	14.23
Melons-Watermelon	3,601.8	30.03

Table 7. Carbon density estimates of different crops using whole tree/plant biomass and typical planting density.

Crop Name	LANDFIRE Existing Vegetation Type (EVT)	Aboveground Carbon (MgC/ha)	Belowground Carbon (MgC/ha)	Source
Thompson grapes	Western Warm Temperate Vineyard	4.13	2.31	Carlisle <i>et al.</i> (2010)
Cabernet Sauvignon	Western Warm Temperate Vineyard	1.88	0.94	Carlisle <i>et al.</i> (2010)
Chenin Blanc	Western Warm Temperate Vineyard	6.01	2.76	Carlisle <i>et al.</i> (2010)
Almond Orchard	Western Warm Temperate Orchard	29.23	ND	DeJong (2013)
Avocado (defruited)	Western Warm Temperate Orchard	11.91	3.37	Rosecrance and Lovatt (2003)
Avocado (heavy fruiting)	Western Warm Temperate Orchard	15.08	5.25	Rosecrance and Lovatt (2003)
Orange	Western Warm Temperate Orchard	17.46	7.17	Morgan <i>et al.</i> (2006)
Alfalfa	Western Warm Temperate Close Grown Crop	14.78	ND	Putnam (2015)

Close Grown Crop EVT, Row Crop EVT, Row Crop-Close Grown Crop EVT

Close grown crop types included Alfalfa, Rice, Oats, and Barley. Biomass and carbon values were weighted based on the statewide acreage allocation of each crop type. A single weighted carbon stock value was then used for the statewide lookup table. Row crops included:

- | | | | |
|------------------------------------|----------------------|-----------------------------------|------------------------------|
| • Tomatoes | • Cotton | • Corn | • Sunflowers |
| • Safflower | • Triticale | • Clover/Wildflowers | • Onions |
| • Double Crop Winter Wheat/Sorghum | • Dry Beans | • Sugar beets | • Potatoes |
| • Misc. Vegetables & Fruits | • Carrots | • Garlic | • Lettuce |
| • Other Crops | • Rye | • Cantaloupe | • Greens |
| • Sorghum | • Watermelons | • Peas | • Broccoli |
| • Pumpkins | • Herbs | • Honeydew Melons | • Sweet Corn |
| • Asparagus | • Peppers | • Double Crop Lettuce/Durum Wheat | • Squash |
| • Mint | • Sweet Potatoes | • Cabbage | • Vetch |
| • Double Crop Lettuce/Cantaloupe | • Canola | • Cauliflower | • Double Crop Lettuce/Cotton |
| • Double Crop Winter Wheat/Cotton | • Sugarcane | • Cucumbers | • Radishes |
| • Pop or Orn Corn | • Other Small Grains | • Double Crop Lettuce/Barley | • Eggplants |

Total above ground yield of crop (for barley, corn, sorghum, sugar beets, cotton, oats, beans, rice, sunflower, wheat and soybean) or peak forage (hay and alfalfa) yield for grazing lands (metric tons biomass/hectare) was needed to calculate above ground C stocks. We used *Equation 1* (below) and Table 8 below (Eve *et al.* 2014, adapted from West *et al.* 2010) to provide a method to convert crop yield to C stocks. The approach was discussed with Mark Easter of the Natural Resource Ecology Laboratory at Colorado State University, who has experience working with similar data and calculations for IPCC reports. Mr. Easter affirmed the approach was appropriate for developing peak herbaceous carbon stock values (Table 9).

Equation 1. The following equation used to calculate aboveground herbaceous biomass carbon stock for harvested crops (adapted from Eve *et al.* 2014 - Equation 3-3).

$$H_{Peak} = \left(Y_{dm} / HI \right) \times C$$

Where:

- H_{Peak} = Annual peak above ground herbaceous (H) biomass carbon stock (metric tons C ha⁻¹ year⁻¹)
- Y_{dm} = Crop harvest or forage yield (Y), corrected for dry matter (dm) content (metric tons C ha⁻¹ year⁻¹); dry matter content of harvested crop biomass or forage is dimensionless and derived from Table 8 below.
- HI = Harvest Index (dimensionless, from Table 8 below)
- C = Carbon fraction of above ground biomass (0.47 gC/g biomass assumed)

Yield (e.g., in bushels per acre) was obtained for each county in California from a query of the National Agricultural Statistics Service (NASS, <http://quickstats.nass.usda.gov/>).

Table 8. Dry matter content factor and harvest Index for common crop types (Summarized from Table 3-5 in Eve *et al.* 2014).

Crop	Dry Matter Content	Harvest Index
Wheat	0.865	0.39
Beans	0.84	0.46
Corn	0.86	0.53
Cotton	0.92	0.40
Oats	0.865	0.52
Rice	0.91	0.42
Hay/alfalfa	0.87	0.95
Sugar beets	0.15	0.40
Sunflower	0.91	0.27

Table 9. Carbon estimates of row and close grown crops by agricultural district.

Agriculture District	District Code	Commodity	Yield (Y)	Dry Matter (DM)	Y(DM)	Harvest Index (HI)	Carbon Content	Herbaceous Peak Carbon (MgC ha⁻¹ yr⁻¹)
Southern California	80	Sugar beets	99.90	0.15	14.985	0.4	0.47	17.61
Other Districts, All Counties	98	Sunflower	1.35	0.91	1.229	0.27	0.47	2.14
Other Districts, All Counties	98	Sunflower	1.30	0.91	1.183	0.27	0.47	2.06
Central Coast	40	Wheat	4.69	0.865	4.057	0.39	0.47	4.89
Northeast	30	Wheat	6.16	0.865	5.332	0.39	0.47	6.43
Other Districts, All Counties	98	Wheat	5.03	0.865	4.347	0.39	0.47	5.24
Sacramento Valley	50	Wheat	5.25	0.865	4.538	0.39	0.47	5.47
San Joaquin Valley	51	Wheat	5.73	0.865	4.955	0.39	0.47	5.97
Siskiyou-Shasta	20	Wheat	6.03	0.865	5.216	0.39	0.47	6.29
Southern California	80	Wheat	6.63	0.865	5.732	0.39	0.47	6.91
Other Districts, All Counties	98	Barley	4.36	0.865	3.774	0.46	0.47	3.86
Sacramento Valley	50	Barley	2.60	0.865	2.251	0.46	0.47	2.30
Other Districts, All Counties	98	Beans	1.67	0.84	1.403	0.46	0.47	1.43
Sacramento Valley	50	Beans	2.02	0.84	1.697	0.46	0.47	1.73
San Joaquin Valley	51	Beans	2.41	0.84	2.024	0.46	0.47	2.07
Other Districts, All Counties	98	Corn	8.35	0.86	7.179	0.53	0.47	6.37
Sacramento Valley	50	Corn	11.26	0.86	9.687	0.53	0.47	8.59
San Joaquin Valley	51	Corn	9.99	0.86	8.588	0.53	0.47	7.62
Other Districts, All Counties	98	Corn	51.75	0.74	38.295	0.95	0.47	18.95
San Joaquin Valley	51	Corn	59.63	0.74	44.123	0.95	0.47	21.83
Sacramento Valley	50	Cotton	1.52	0.92	1.398	0.4	0.47	1.64
San Joaquin Valley	51	Cotton	1.56	0.92	1.433	0.4	0.47	1.68
San Joaquin Valley	51	Cotton	1.80	0.92	1.652	0.4	0.47	1.94
Southern California	80	Cotton	2.07	0.92	1.905	0.4	0.47	2.24
Other Districts, All Counties	98	Oats	3.58	0.865	3.098	0.52	0.47	2.80
Siskiyou-Shasta	20	Oats	4.50	0.865	3.893	0.52	0.47	3.52
Sacramento Valley	50	Rice	8.61	0.91	7.835	0.42	0.47	8.77
San Joaquin Valley	51	Rice	7.79	0.91	7.089	0.42	0.47	7.93
Sierra Mountains	60	Rice	7.90	0.91	7.189	0.42	0.47	8.04
California	N/A	Hay	12.08	0.87	10.512	0.95	0.47	5.20
California	N/A	Hay/Alfalfa	14.63	0.87	12.724	0.95	0.47	6.29

Agricultural vegetable and fruit crops such as tomatoes and broccoli are problematic for developing herbaceous carbon stock data based on yield information because significant biomass may be left on site following harvesting (i.e., post-harvest residue). However, we found information on a few crops for which the residue volume has been quantified (Mitchell *et al.* 1999) in a way that allows a calculation of post-harvest residue carbon. In cases where post-harvest residue data were available, we combined the yield data (from NASS) with residue data to calculate peak crop biomass carbon using Equation 2 (See also Table 9).

Equation 2. Equation used to calculate aboveground peak crop biomass carbon.

$$Crop_Biomass_{Peak} = Herbaceous_Residue_{Peak} + Vegetable_Fruit_{Yield}$$

We found there was not sufficient aboveground biomass data on numerous agricultural crops in California. Table 10 lists crops with carbon estimates for the total yield only. Further communication with Mark Easter (Colorado State University) and Dr. Holly Gibbs (University of Wisconsin) confirms that much of this data does not exist, thus default values or similar crop carbon stock values must be used in the assessment.

Table 10. Carbon estimates of crops using harvest residue biomass estimates (Mitchell *et al.* 1999) and yield estimates (National Agricultural Statistics Service 2015).

Commodity	Aboveground Carbon Density (MgC/ha)
Corn	5.04
Broccoli	3.45
Cotton	2.75
Wheat	2.54
Sugar beet	2.02
Safflower	1.42
Tomato	1.52
Lettuce	1.03
Garlic	0.49
Onion	0.30

Developed Ruderal Grassland EVT

California coastal and valley grasslands have published data estimating above and below ground carbon stocks (Ryals and Silver 2013), and Li *et al.* (2012) used MODIS satellite imagery data to calculate estimates of Net Primary Productivity (NPP) for California rangelands, from which carbon estimates were derived.

Urban Deciduous, Urban Evergreen, Urban Mixed

We obtained urban forest carbon estimates by using biomass stock data (tons/acre) from Bjorkman *et al.* (2015). These data were summarized by county for California (Table 11). A simple conversion was performed to convert to Mg/hectare and the standard carbon coefficient of 0.47 grams carbon/gram (gC/g) biomass was applied.

Table 11. Urban forest carbon estimates by county in California (from Bjorkman *et al.* 2015).

County	Biomass (tons)	Acres	Biomass (tons/acre)	Tons C/acre	MgC/ha
Alameda	1,548,926	174,989	8.85	4.16	9.36
Amador	68,830	4,934	13.95	6.56	14.75
Butte	1,120,363	54,115	20.70	9.73	21.89
Calaveras	96,924	6,637	14.60	6.86	15.44
Colusa	21,680	3,165	6.85	3.22	7.24
Contra Costa	2,311,140	196,645	11.75	5.52	12.43
Del Norte	249,866	7,641	32.70	15.37	34.58
El Dorado	1,206,225	48,422	24.91	11.71	26.34
Fresno	1,070,018	136,945	7.81	3.67	8.26
Glenn	28,196	5,408	5.21	2.45	5.51
Humboldt	675,532	30,220	22.35	10.51	23.64
Imperial	23,573	27,228	0.87	0.41	0.92
Inyo	30,132	2,739	11.00	5.17	11.63
Kern	532,475	141,401	3.77	1.77	3.98
Kings	97,818	25,230	3.88	1.82	4.10
Lake	153,909	17,232	8.93	4.20	9.45
Lassen	16,358	3,431	4.77	2.24	5.04
Los Angeles	4,901,846	921,840	5.32	2.50	5.62
Madera	156,203	25,345	6.16	2.90	6.52
Marin	1,810,810	54,653	33.13	15.57	35.04
Mendocino	390,655	18,769	20.81	9.78	22.01
Merced	236,350	44,853	5.27	2.48	5.57
Modoc	5,410	1,223	4.42	2.08	4.68
Mono	26,196	2,127	12.32	5.79	13.02
Monterey	934,368	68,646	13.61	6.40	14.39
Napa	444,623	26,305	16.90	7.94	17.87
Nevada	848,774	30,578	27.76	13.05	29.35
Orange	2,250,163	339,919	6.62	3.11	7.00
Placer	1,551,412	91,290	16.99	7.99	17.97
Plumas	13,904	2,356	5.90	2.77	6.24
Riverside	1,114,534	456,930	2.44	1.15	2.58
Sacramento	1,847,149	213,190	8.66	4.07	9.16
San Benito	22,791	7,324	3.11	1.46	3.29
San Bernardino	1,180,171	403,731	2.92	1.37	3.09
San Diego	3,656,029	504,835	7.24	3.40	7.66
San Francisco	397,782	30,318	13.12	6.17	13.87
San Joaquin	481,933	101,226	4.76	2.24	5.03
San Luis Obispo	744,037	62,726	11.86	5.57	12.54
San Mateo	1,617,695	91,160	17.75	8.34	18.77
Santa Barbara	651,121	68,116	9.56	4.49	10.11

County	Biomass (tons)	Acres	Biomass (tons/acre)	Tons C/acre	MgC/ha
Santa Clara	2,351,202	211,971	11.09	5.21	11.73
Santa Cruz	1,356,195	51,052	26.57	12.49	28.09
Shasta	570,789	49,843	11.45	5.38	12.11
Sierra	21	5	4.62	2.17	4.88
Siskiyou	93,922	7,860	11.95	5.62	12.64
Solano	532,515	73,643	7.23	3.40	7.65
Sonoma	1,450,695	92,505	15.68	7.37	16.58
Stanislaus	550,427	76,754	7.17	3.37	7.58
Sutter	129,849	15,808	8.21	3.86	8.69
Tehama	79,142	10,568	7.49	3.52	7.92
Tulare	325,787	71,885	4.53	2.13	4.79
Tuolumne	743,715	20,108	36.99	17.38	39.11
Ventura	703,152	143,916	4.89	2.30	5.17
Yolo	265,102	30,487	8.70	4.09	9.20
Yuba	67,342	11,986	5.62	2.64	5.94

Default IPCC Values

We were unable to obtain appropriate carbon stock values for several LANDFIRE EVT's from literature or calculated from available data. In those cases, we used IPCC default values (Table 12). IPCC default values were obtained by using the value of 5 Mg/ha for cultivated and managed land and 1 Mg/ha for bare/fallow/idle areas (Ruesch and Gibbs 2008). Default values for developed ruderal coniferous/deciduous/mixed forests were derived from Penman *et al.* 2003 using dry matter values for temperate forests ≤ 20 years old (multiplied by 0.47gC/g biomass).

Table 12. Default IPCC Values for carbon for cultivated and managed land, bare areas, and water, snow, ice and artificial surfaces. IPCC default values of 5 were used for cultivated and managed land and 1 for bare/fallow/idle areas.

GLC2000 Class	FAO Ecofloristic Zone, and Continental Region, and Frontier Class	Carbon Value
16: Cultivated and Managed Land	All	5.0
19: Bare Areas		1.0
20-23: Water, Snow and Ice, and Artificial Surfaces		0.0

Source: Ruesch and Gibbs 2008.

Step 5: Evaluation of Dead Carbon Pools

As a potential refinement to Battles *et al.* (2014) GHG Inventory Tool, we quantified the differences in dead carbon pools when estimated using FIA field data and fuel loading plot data from various study sites, versus when estimated from LANDFIRE's FCCS and FBFM (Scott and Burgan fire behavior fuel model) mapping products. Understanding differences could help to validate assumptions and improve estimates in dead wood carbon pools. Results of this evaluation could likewise be used to inform procedures for statewide carbon pool inventory and stock change assessment.

Field Plot Data Analysis

This comparison utilized field data from 1,697 mixed conifer plots at five locations in the Sierra Nevada. Three of the field locations were part of the Sierra Nevada Adaptive Management Project (SNAMP; including “Last Chance” [on the Tahoe National Forest, and portions of the Eldorado National Forest and about 22 km NE of Forest Hill], “Sugar Pine”, and “Cedar Valley” [both NE about 9 to 13 km of Oakhurst, CA in the Sierra National Forest]) study (e.g., Collins *et al.* 2011) and measured in 2007 and 2008, one was from the Sagehen Experimental Forest and measured in 2005, and the other from the Blodgett Experimental Forest site of the Fire and Fire Surrogates Study which was measured in 2003. All plots were untreated and unburned except the Blodgett plots whose fuels treatments were reflected in the 2008 LANDFIRE disturbance mapping.

All field data was compared to Scott-Burgan fire behavior fuel models (FBFM40) and Fuel Characteristic Classification System (FCCS) fuelbeds attributed in LANDFIRE 1.1.0 (2008). The Fuel Characteristic Classification System calculates and classifies fuelbed characteristics and their potential fire behavior. FCCS fuelbeds represent fuels throughout much of North America and were compiled by LANDFIRE from published literature, fuels photo series, other fuels data sets and expert opinion. FCCS fuelbeds have been mapped in LANDFIRE and are preloaded in the USFS Fuel and Fire Tools application. Similarly, LANDFIRE mapped Scott and Burgan fire behavior fuel models across the nation. These fuel models are designed to work with the Rothermel (1983) fire spread model and are defined to produce certain characteristic fire behavior. Unlike the FCCS fuelbeds, they were not intended for comprehensively describing live and dead fuel loads and therefore contain less information than the fuelbed definitions. However, each Scott and Burgan fuel model includes in its definition masses of dead fuel specified by size class that may be converted into carbon.

The dead carbon pools that were used for this comparison included:

- Duff fuel load
- Litter fuel load
- Duff + litter fuel load
- 1-hr fuel load
- 10-hr fuel load
- 100-hr fuel load
- 1000-hr sound fuel load
- 1000-hr rotten fuel load
- Total 1000-hr fuel load
- Total surface fuel load.

The duff + litter fuel load is measured together in the field and is not simply the sum of the two components. While each FCCS fuelbed represents all of the pools listed above, the FBFM40 only included 1) 1-hr, 2) 10-hr, and 3) 100-hr fuel loads.

We obtained spatial coordinates for each field plot from various study sites and assigned those coordinates to associated LANDFIRE data layer pixels for FBFM40 and FCCS. We then compared the field plot data values and LANDFIRE values for each of the following dead carbon pools. The comparison was evaluated as $(1 - (\text{field value} - \text{LANDFIRE value}) / \text{field value}) \times 100$. The result indicated how favorably the LANDFIRE values compare to the field value with 100% being an exact match, greater than 100% showing that LANDFIRE over-predicted relative to the field measure, and less than 100% showing that LANDFIRE under-predicted relative to the field

measure. We averaged this calculated field measure within each dead carbon pool at each site. At some plots nothing was actually measured due to an absence of a dead carbon pool and were excluded to avoid a divided by zero error in the calculations. Nothing was wrong with the plots where a zero value was measured however they produced a division by zero error in the calculation described above. We also averaged relative differences for each carbon pool across all five sites. All comparisons were performed in tonnes of biomass per hectare for both field data and LANDFIRE data.

Results within each pool were fairly consistent regardless of site (Tables 13 and 14). Using the comparison calculation described above the best matches relative to field value averaged across all sites were:

1. FCCS litter (62%)
2. FCCS 1000-hr sound (65%)
3. FCCS 1000-hr rotten (121%)
4. FCCS 1-1000-hr sound + rotten (121%)
5. FCCS 100-hr (188%)

The worst matches relative to field value averaged across all sites were:

1. FBFM40 1-hr (4047%)
2. FBFM40 total (1885%)
3. FCCS 1-hr (782%)
4. FBFM40 10-hr (725%)
5. FCCS total (651%)

Tables 13 and 14 below list the results of the comparison at each field data location below.

Table 13. Percent difference in LANDFIRE FCCS dead carbon pool relative to field plot data (S = sounds, R = rotten).

Site	Duff	Litter	Duff + Litter	1-hr	10-hr	100-hr	1000-hr (S)	1000-hr (R)	1-1000 (S+R)	Total
SNAMP-Last Chance	370%	38%	140%	1,151%	759%	203%	65%	199%	1,817%	357%
SNAMP-Sugar Pine	153%	25%	146%	1,234%	697%	154%	43%	107%	3,640%	678%
SNAMP-Cedar Valley	195%	40%	76%	820%	612%	204%	87%	121%	1,071%	192%
Sagehen	474%	57%	219%	568%	328%	170%	41%	89%	4,534%	431%
FFS-Blodgett	309%	59%	253%	293%	216%	129%	81%	67%	871%	1361%
Average (n=1,697 plots)	300%	44%	167%	813%	522%	172%	63%	117%	2,387%	604%
Average (weighted by number of plots)	407%	62%	224%	782%	514%	188%	65%	121%	1218%	651%

Table 14. Percent difference of LANDFIRE FBFM40 dead carbon pool relative to field data.

Site	1-hr	10-hr	100-hr	Total
SNAMP-Last Chance	4,624%	854%	269%	1,661%
SNAMP-Sugar Pine	4,724%	806%	233%	2,633%
SNAMP-Cedar Valley	3,868%	955%	389%	2,458%
Sagehen	4,782%	701%	312%	2,187%
FFS-Blodgett	1,673%	407%	85%	937%
Average (1,697 plots)	3,934%	745%	258%	1,975%
Average (weighted by # of plots)	4,047%	725%	249%	1,885%

FIA Data Comparison

Next we compared 2001 and 2008 LANDFIRE FCCS and FBFM40 products to FIA plot data. We used the most recent FIA database ([FIADB 1.6.0.02](#), 2015-05-08) for California. The carbon pools of interest in this database came from field measurements conducted from 2001 to 2010. We used this database to summarize carbon in 1-hour, 10-hour, 100-hour, litter, and duff fuel loads within forestland in each of the state's ecological subregions. We also computed zonal statistics using ArcMap to find the majority value of 2001 and 2008 LANDFIRE FCCS fuelbed and FBFM within each of the same ecological subregions. We were then able to compare carbon pools from FIA data to LANDFIRE carbon pools across the state.

There was little difference between the 2001 and 2008 versions of LANDFIRE's FCCS and FBFM layers when compared to the most current FIA data within forestland and summarized by [ecological subregion](#) (McNab *et al.* 2005). The 1-hr FCCS fuelbed values tended to be close to 700% greater than the corresponding FIA data while the 10-hr dead carbon pools were about 220% greater. The FCCS 100-hr fuels were about 85% of the corresponding FIA values. Litter and duff were approximately 21% and 240%, respectively, of the FIA data. The FBFM dead carbon pools, on the other hand, differed more substantially from the FIA data. The FBFM 1-hr dead carbon pools were more than 3,000% greater than the FIA data while the FBFM 10-hr dead carbon pools were more than 300% greater than the FIA data. The FBFM 100-hr dead carbon pools were about 60% of the FIA values. Overall, the FCCS dead carbon pools provided a better fit with FIA data across the state's ecological subregions (Tables 15).

Table 15. Percent difference across all ecological subregions between FCCS and FIA, and FBFM and FIA

Dataset	FCCS relative to FIA					FBFM relative to FIA		
	1-hr	10-hr	100-hr	litter	duff	1-hr	10-hr	100-hr
2008 LANDFIRE	690%	222%	86%	21%	240%	3,110%	335%	58%
2001 LANDFIRE	668%	217%	82%	24%	231%	3,348%	379%	60%

The results of this analysis support the use of FCCS fuel beds to estimate carbon in aboveground dead wood carbon pools. However, additional work is needed to develop a multiplier or conversion factor that could be used to better align different FCCS fuel beds with field and/or FIA data.

Step 6: Identify Carbon Considerations of Forest Management and Harvested Wood Products

Quantifying carbon stock changes associated with direct human-induced ‘degradation’ of forests and ‘devegetation’ of other vegetation types is complicated by variations in the intensity of activities such as timber harvesting (including both commercial and non-commercial operations) according to ownership type. Vegetation management and harvest activities lead to carbon stock changes that are difficult to quantify using remotely-sensed data since such activities are periodic in nature and harvested carbon stocks can recover at varying rates in between data acquisition years. Assessing stock changes from vegetation management and harvest activities requires calibration between site-level removal (harvest) data and remotely-sensed data that is adjusted for land ownership type and the temporal lag of monitoring data. Moreover, removed forest biomass should not be accounted for as an immediate emission since a fraction will be sequestered in wood products over a variable lifetime.

In response to this challenge, we developed: 1) methodologies that use existing datasets to allocate areas with harvest activities, and categorize these harvest activities to the largest extent possible with biomass removal intensities, and 2) a crosswalk from defined harvest activities towards 100-year lifecycle emissions (losses) associated with these harvest activities and the wood products derived from these activities.

Harvest Operations in California

We applied the LANDFIRE “Disturbance” 1999-2012 data layer to California state boundaries and filtered for all harvest related disturbance types, namely ‘Clearcut’, ‘Thinning’, and ‘Harvest’ (Table 16). The following paragraph describes the data layer (LANDFIRE 2016) as:

LANDFIRE disturbance data are developed to provide temporal and spatial information related to landscape change for determining vegetation transitions over time and for making subsequent updates to LANDFIRE vegetation, fuel and other data. Disturbance data include attributes associated with disturbance year, type, and severity. These data are developed through use of Landsat satellite imagery, local agency derived disturbance polygons, and other ancillary data. From the abstract: The disturbance data are developed through a multistep process. Inputs to this process include; Landsat imagery and derived NBR (normalized burn ratio) data; polygon data developed by local agencies for the LF Events geodatabase effort; fire data obtained from MTBS (Monitoring Trends in Burn Severity), BARC (Burned Area Reflectance Classification), and RAVG (Rapid Assessment of Vegetation Condition after Wildfire) fire mapping efforts, PAD (Protected Area Database) data, and Smartfire ignition point buffer polygons (buffer distance dependent on sensor accuracy). LANDSAT imagery and derived NBR data are not included in Alaska disturbance grid development. LF Event polygon data are provided to LANDFIRE by various local, regional, and national agencies and organizations. Disturbance type and year information is included as attributes for each polygon and transferred to the disturbance grids. Severity is determined by using dNBR (difference Normalized Burn Ratio) data classified into high, medium, and low severity levels based on dNBR standard deviation thresholds. Vegetation Change Tracker (VCT) algorithms (Huang, et. al. 2008) were used to identify disturbances outside of LF Events for the LF2008 effort (years 1999-2008). Multi-Index Integrated Change Algorithm (MIICA) methods (Jin, et. al. 2013) were used to identify additional change in 2008 as well as disturbances in 2009 and 2010 for the LF2010 effort. Since disturbance type (i.e. causality) is not determined in the VCT or MIICA processes, a spatial analysis is done comparing the output to buffered (500 meter)

LF Events, Protected Area Database GAP Status information (land use and management characteristics), and Smartfire ignition point buffer polygons. While not providing a precise type of disturbance, this analysis provides information useful for narrowing down the types of disturbance that could or could not typically occur. Each zone has 13 disturbance grids, one for each year 1999 to 2012. Each grid is attributed with year, disturbance type (if known, otherwise a description of possible types), severity, data sources, and confidence (type and severity). VdistYEAR grids are a composite of the last ten years of disturbance grids recoded by disturbance type, disturbance severity, and time since disturbance YEAR to meet LANDFIRE vegetation transition modeling needs. Fire occurrences take precedence, followed by the most recent disturbance taking precedence.

Table 16. Harvest-related disturbance types in LANDFIRE Disturbance 1999-2012 dataset (Source: LANDFIRE 2015).

Attribute	Enumerated Value	Enumerated Value Description
Dist_Type	Clearcut	The cutting of essentially all trees, producing a fully exposed microclimate for the development of a new age class.
Dist_Type	Harvest	A general term for the cutting, felling, and gathering of forest timber. The term harvest was assigned to events where there was not enough information available to call them one of the 2 distinct types, clearcut or thinning.
Dist_Type	Thinning	A tree removal practice that reduces tree density and competition between trees in a stand. Thinning concentrates growth on fewer, high-quality trees, provides periodic income, and generally enhances tree vigor.

Wood Products Carbon Assessment

We identified the percentage of merchantable timber volume from the total study area landscape (i.e., California) to estimate carbon loss for a given harvest activity (i.e., each for “clearcut”, “harvest”, and “thinning” from Table 16 above). Using measured data from 28 harvest sites (partial and clearcut) covering a total of 2,781 ha (Stewart and Nakamura 2012), and we generated average carbon loss from each harvest activity type, as well as multipliers on carbon stored in logs for different harvest operations on private timberlands (Table 17).

Table 17. Landscape carbon loss and merchantable volumes on private (Stewart and Nakamura 2012) and public (Saah *et al.* 2012) land.

Ownership Type	Harvest Type	Mean Total Harvest Carbon Density (MgC/ha)	Mean Merchantable Carbon Density (MgC/ha)	Percent Merchantable of Total Harvest Carbon Density
Private	Clearcut	48.9	43.4	89%
Public	Clearcut	48.9	43.4	89%
Private	Partial cut	21.0	7.3	42%
Public	Partial cut	11.8	6.9	42%
Private	Harvest	N/A	N/A	72%
Public	Harvest	N/A	N/A	44%

For carbon loss (emissions) and harvest carbon density estimates on public lands, we compared timberland carbon densities for mature forest stands on public and private ownerships across California using Forest Inventory Analysis data (FIA 2015) for the time period of interest (1999 to 2012). We detected no discernable difference in carbon density (MgC/ha) and therefore assumed the same carbon stocking (48.9 MgC/ha) and wood product carbon density (43.4 MgC/ha) from public timberlands as from private timberlands.² For thinning/partial cuts on public lands, we used estimates of 6.9 MgC/ha in merchantable carbon densities (Saah *et al.* 2012) and assumed a similar ratio in merchantable vs. total harvest carbon density as for private lands. To convert MgC to million board foot (mmbf), we used a conversion rate of 572 MgC/mmbf (Skog and Nicholson 2000).

For the LANDFIRE disturbance “harvest” category, which enumerates all harvest sites that could not be allocated to either thinning/partial cut or clearcut activities, we calculated a merchantable vs. total harvest volume ratio based on the normalized total harvest volumes reported for LANDFIRE disturbance categories for “thinning” and “clearcut” and multiplied by the respective ratio of merchantable volumes for each of these two categories.³

Next, we generated carbon loss multipliers over a 100 year timeframe for those harvest volumes. Since wood products will store carbon in-use and post-use when landfilled the fraction of carbon stored in these wood products over a given timespan needs to be subtracted from landscape carbon loss.

Using numbers from Smith *et al.* (2006), which are the basis for the national 1605(B) [Voluntary Reporting of Greenhouse Gases Program](#) lookup tables as well as estimates from the University of California (2015)⁴ Carbon Sequestration Tool for THPs, we estimated that 36% or 46% of C, respectively, would be permanently stored in wood products over a 100 year time frame under a normal California wood products life span. Carbon not used in wood products (unrecovered residues, forest and sawmill residues used for bioenergy, etc.) was assumed to be emitted completely over this timeframe.

Validation

Validation of LANDFIRE outcomes were based on: 1) harvested acreage on private timberlands as reported by CALFIRE (2010), 2) carbon stock loss estimates on a per acre basis by harvest type using various other references (Table 17), as well as 3) reported merchantable volume estimates as reported by the California State Board of Equalization (BOE 2015, Table 18). Using this data, the BOE Timber Yield Tax program sets the harvest value of timber and collects an in lieu tax when it is harvested. Not all carbon loss associated with harvest activities in the first as well as last year of the time period of interest 2001 to 2010 were captured by the LANDFIRE dataset due to continuous data collection efforts. We therefore included only 50% of the BOE reported harvest volumes for the first (2001) and last (2010) year.

Landscape Harvest Impact 2001-2010

There is no spatially explicit dataset available to validate acreage outcomes from the LANDFIRE Disturbance layer except for CALFIRE data on harvests from private lands from 2001 to 2008 (CALFIRE 2010) totaling 395,611 ha. Prorating this acreage to 2010 results in a total acreage

² See ‘Wood product C pools from CA 1999-2012 2015-10-16.xls’; sheet ‘FIA owner C density’.

³ See ‘Disturbance_2001_2010 2015-11-16.xls’; sheet ‘Dashboard’; cell B28/C28.

⁴ Numbers for mixed conifer stands, see also ‘Wood product C pools from CA 1999-2012 2015-10-16.xls’; sheet ‘Dashboard’, cell J8.

estimate of 494,513 ha which is reasonably close to the acreage reported for private lands in the 2001-2010 LANDFIRE Disturbance layer totaling 432,283 ha (Table 19). While the difference for clearcut acreage is marginal (83,446 vs 95,474 for LANDFIRE data vs prorated CALFIRE data, respectively), most of the difference is grounded in accounting for the correct acreage for partial harvests. This comparison suggests that most of what has been reported as uncategorized harvest in the LANDFIRE Disturbance layer is most likely partial harvest. Partial harvests are much more diverse in nature⁵ than clearcuts and are not as easily discernable as clearcuts. The significance of partial harvests with lower merchantable volumes per acre also explain the fact that while public lands account for only 10 % of harvested merchantable volume (Table 18), they also account for a total of 29% of total carbon loss through harvests (Table 19).

Table 18. Harvested merchantable volumes in million board feet of timber (mmbf) in California 2001-2010 (BOE 2015).

Year	Private	Public	Total	Public as % of Total
2001	843,700	73,216	916,916	8%
2002	870,012	96,668	966,680	10%
2003	862,576	88,660	951,236	9%
2004	911,196	64,636	975,832	7%
2005	855,140	131,560	986,700	13%
2006	818,532	114,400	932,932	12%
2007	823,108	106,964	930,072	12%
2008	728,156	56,628	784,784	7%
2009	426,140	34,320	460,460	7%
2010	586,300	77,792	664,092	12%
Total	7,724,860	844,844	8,569,704	10%

Table 19. Aboveground live carbon loss by harvest type in Mg C/ha 2001-2010.

LANDFIRE Harvest type	Private			Public		
	Net C loss (Mg C)	Area (ha)	Net C loss (Mg C/ha)	Net C loss (Mg C)	Area (ha)	Net C loss (Mg C/ha)
Clearcut	3,110,624	83,446	37	291,798	7,594	38
Harvest	2,477,681	186,152	13	608,990	41,139	15
Thinning	1,739,266	162,685	11	2,132,683	138,337	15
Total	7,327,571	432,283	17	3,033,471	187,070	16

Total net C emissions from harvest activities including carbon stored long-term in wood products were 5,576,968 to 6,682,872 Mg C (Table 20) when using University of California (2015) or Smith *et al.* (2006) wood products carbon coefficients, respectively. Post-use assumptions have a high impact on these numbers, most notably if wood waste is land filled or incinerated. It is safe to assume that most of wood products within California origin are eventually landfilled since less than 1% of the in-state log production is exported to other countries (McIver et al. 2015) where waste incineration might be more applicable than landfilling.

⁵ See e.g. CALFIRE partial harvest categories Commercial Thin; Fuelbreak/Defensible Space; Group Selection; Rehabilitation; Right of way (Road Construction); Sanitation-Salvage; Seed Tree Removal; Seed Tree Seed Step; Selection

Table 20. Total net C emissions in Mg for 2001-2010 including carbon stored in wood products for >100 years.

	Smith et al. (2006)			University of California (2015)		
	Private	Public	Total	Private	Public	Total
Clearcut	(2,006,353)	(188,209)	(2,194,562)	(1,674,335)	(157,064)	(1,831,399)
Harvest	(1,598,104)	(392,799)	(1,990,903)	(1,333,644)	(327,797)	(1,661,442)
Thinning	(1,121,826)	(1,375,581)	(2,497,407)	(936,183)	(1,147,945)	(2,084,128)
Total	(4,726,283)	(1,956,589)	(6,682,872)	(3,944,162)	(1,632,806)	(5,576,968)

Reported Harvest Intensities

Using LANDFIRE data, we calculated average aboveground live carbon loss of 11 to 15 Mg C/ha for thinning harvests and 37 to 38 Mg C/ha for clearcuts depending on ownership type (Table 19). While validated for thinning operations (11.8 to 21.0 Mg C/ha; Table 17), these harvest intensities only partly support other data points for clearcuts (48.9 Mg C/ha in Stewart and Nakamura 2012) which tend to be slightly higher (Table 17).

Reported Harvest Volumes

Total harvest volume was calculated to be 6,691,256 Mg C which accounts for 86% of the BOE reported harvest volume from mid-2001 to mid-2010. The remaining difference is rooted in a variety of factors including unaccounted in-growth on harvest sites in the LANDFIRE dataset. On an interesting side note, the LANDFIRE approach suggests a higher volume in merchantable volumes provided from public timberlands (21% of total; Table 21) vs BOE reported numbers (10% of total; Table 18) when converting LANDFIRE values (Mg C) back to board feet. While BOE receives its numbers on merchantable harvest volumes directly from timber receipts, the LANDFIRE Disturbance layer receives its data from multiple sources, frequently relying on indirect assumptions on harvested volumes and generalized multipliers converting carbon to board feet.

Table 21. Merchantable volumes in mmbf and % of total harvested 6/2001 to 6/2010 based on LANDFIRE data.

	Private		Public		Total	
Clearcut	4,827	41%	453	4%	5,280	45%
Harvest	3,118	27%	466	4%	3,585	31%
Thinning	1,277	11%	1,557	13%	2,834	24%
Total	9,222	79%	2,476	21%	11,698	100%

Results suggest that the GHG Inventory Tool accounts reasonably well for harvested wood products. While total acreage affected by harvest as well as harvested merchantable volume activity is generally supported by other data especially for clearcuts, harvest intensities are only partly supported by other data points which seem to be higher for clearcuts and lower for thinning operations.

Step 7: Accounting for Undetected Biomass Growth

The ordinal nature of the LANDFIRE height (EVH) and cover (EVC) variables may lead to underestimation by our methods of carbon changes in pixels that experience no change in vegetation type (EVT). For fractional cover, LANDFIRE defines ten classes that increase in even steps of 10%. For tree height, LANDFIRE classes step up more steeply as height increases. If the average height or cover of a pixel changes but does not cross into the next class, our method records no change (positive or negative) in carbon density. Because growth can occur slowly relative to the nine-year period of our analysis, our methods can underestimate carbon changes due to growth within a cover or height class. Consequently, our stock-change assessment may not completely capture growth as immediately as land cover change.

Recently released data from FIA plots that the Forest Service has resampled over the last decade allow us to estimate the magnitude of our potential underestimate of growth in tree-dominated vegetation. We calculated the plot-level biomass of the 966 plots in California (all tree-dominated) measured in 2001 and 2002 and re-measured 10 years later (FIA database version 6.0, October 2, 2014). The distribution of plots that added biomass was different from the plots that lost biomass (Figure 4). Out of the 966 plots, 274 plots lost biomass (range: -0.03 to -428 Mg ha⁻¹). In contrast, there were many more small gainers; 686 plots gained biomass (range: 0.04 to 202 Mg ha⁻¹, Figure 5). Given the way small change are detected (i.e., changes in cover and/or height), we detect large changes better than small changes. Since the preponderance of small changes tend to be gains, we likely underestimate growth.

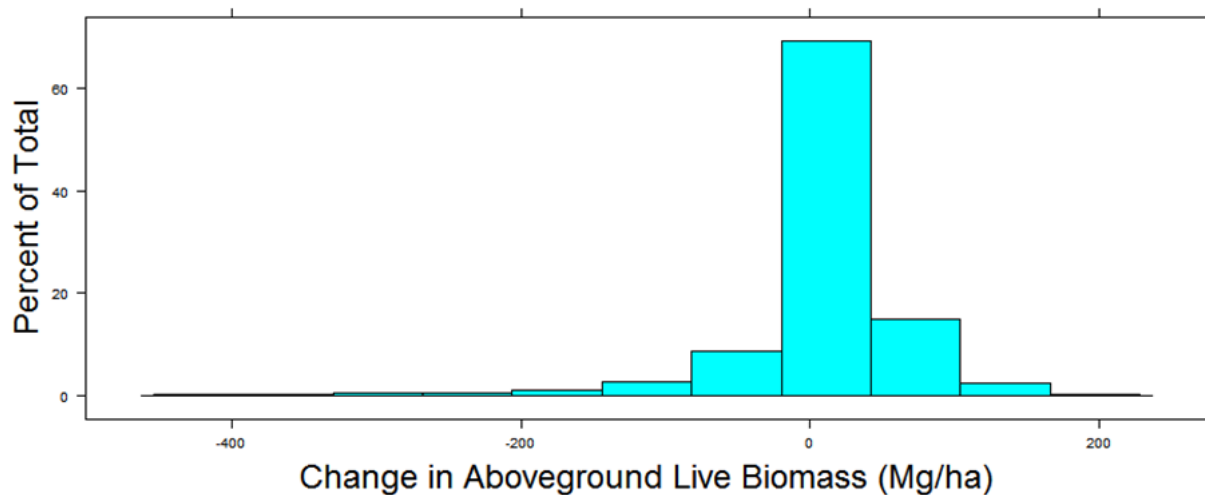


Figure 4. Distribution of changes in aboveground live biomass for 966 repeat measures FIA plots that remained forest from 2001-2002 to their re-measurement dates in 2011 and 2002.

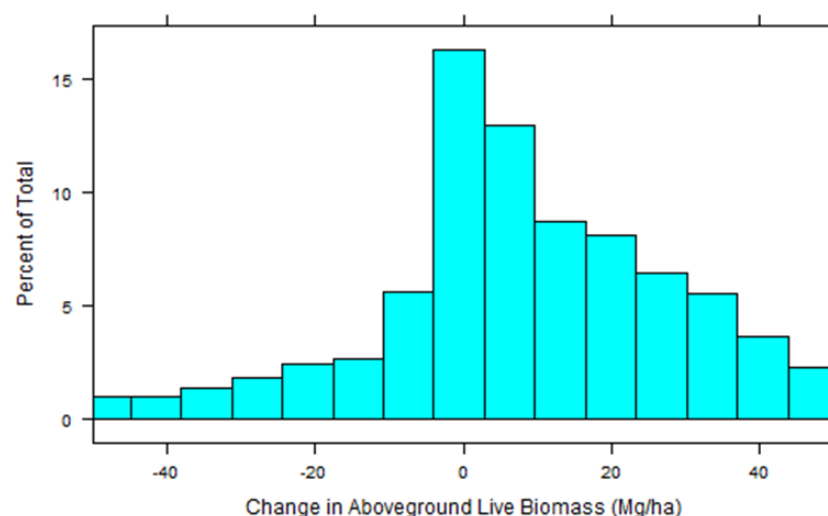


Figure 5. Closer examination of the distribution of changes in aboveground live biomass growth for 966 re-measured plots in the FIA data. This histogram clearly shows the many more small gains in biomass compared to small losses.

To correct this potential bias, we used the results from the re-measured plots. Plot-level aboveground biomass increased 6% over the decade. If those plots comprise a representative

sample and if, in the worst case, our method did not capture any growth, the growth in tree-dominated vegetation types remaining tree-dominated would be underestimated by 0.6% yr⁻¹. In this assessment of stock changes, we corrected for this potential underestimate.

Step 8: Updated Lookup Tables and Geographic Information System Data (the Updated GHG Inventory Tool)

Battles *et al.* (2014) statewide inventory of GHG and associated stock changes assessment hinges on the accuracy of categories and biomass and carbon values represented in the GHG Inventory Tool “biomass classes” lookup table. For each combination of vegetation type, height and cover class, and iteration of LANDFIRE data products (2001, 2008 and 2010), the lookup table developed for this project contains above ground biomass and associated carbon estimates derived from various sources as described in the previous steps. Specifically, a core lookup table, ‘BATTLES_Biomass-LUT_01-08-10_20151029’ serves to link other lookup tables in the updated GHG Inventory Tool and ‘ARB_C_LUT_v2.7.accdb’ database (lookup table database). The lookup table database (ARB_C_LUT_v2.7.accdb) is used to organize all lookup tables in the GHG Inventory Tool. The lookup table database includes the following tables, queries and functions (i.e., macros):

Tables

- **BATTLES_Biomass-LUT_01-08-10_20151029** – is a biomass lookup table that combines existing LANDFIRE vegetation types, height, and cover across 2001, 2008, and 2010. Includes attributes from IPCC landuse, and biomass and carbon values for each row.
- **GUNN_AG_LUTv4** – is a lookup table for agriculture associated LANDFIRE EVTs across 2001, 2008, and 2010. Include carbon density estimates (MTC/ha) for each EVT.
- **GUNN_Urban_CNTY_LUTv4** – is a lookup table that provides carbon densities (MTC/Ha) for urban landuses by county and the source of carbon density estimates.
- **GUNN_Urban_LUTv4** - is a lookup table that provides carbon densities (MTC/Ha) for urban landuses by LANDFIRE EVT.
- **LFc_2001** – is a lookup table that contains combined attributes from LANDFIRE existing vegetation type, height and cover data products from 2001.
- **LFc_2008** - is a lookup table that contains combined attributes from LANDFIRE existing vegetation type, height and cover data products from 2008.
- **LFc_2010** - Table contains combined attributes from LANDFIRE existing vegetation type, height and cover data products from 2010.
- **LUT_Disturbance** – Is a lookup table that code different types of disturbance or timber management activity. Timber management activities are coded starting with a ‘3’ (e.g., a clearcut is coded as ‘35’), fire is coded starting with a ‘2’ (e.g., wildfire is coded as ‘22’), and other disturbance is coded starting with a ‘1’ (e.g., development is coded as ‘13’).
- **LUT_IPCC_ACTIVITY** – is a lookup table that codes landuse conversions using IPCC categories.
- **LUT_IPCC_CODEV1** – is a lookup table that codes IPCC landuse types.
- **LUT_LFyearv1** – is a lookup table that codes LANDFIRE data product years (2001, 2008, and 2010)
- **LUT_ORDER_GROWTHv1** – is a lookup table that codes the LANDFIRE ‘Order’ tier. It is used to assign growth to tree dominated EVTs in each LANDFIRE data product year (2001, 2008, and 2010).

Queries

- IPCC_AC_CODE – this was originally a lookup that is now deprecated.
- ARB_BIOMASS_LUT_v2.7 – is a lookup table that combine urban and agriculture carbon densities with the BATTLES_Biomass-LUT_01-08-10_20151029

Macros

- LUT_EXPORT – converts 'ARB_C_LUT_v2.7.accdb' access database tables and queries into individual excel spreadsheets. These spreadsheets are used for geodatabases to show distribution of biomass and carbon values across California and calculating carbon stock change across LANDFIRE data product years. The following spreadsheets are generated with this macro:
 - LUT_ARB_BIOMASS_v2_7.xls
 - LUT_GUNN_Urban_CNTYv4.xls
 - LUT_IPCC_ACTIVITY.xls
 - LUT_IPCC_CODEv1.xls
 - LUT_LFYEARv1.xls
 - LUT_LUT_Disturbance.xls
 - LUT_ORDER_GROWTHv1.xls

The spreadsheets were then ingested into one to many different geodatabases through ArcGIS model builder. The lookup tables are used for all carbon stock calculations. Geodatabases and associated models for calculating stock changes are organized in the "02_Calculations" folder using the same file structure as the LandCarbon Models used for conducting the calculations contained in the 'Toolbox.tbx' file. Below is a brief description of each step where the model is labeled in sequential order and their results are filed in a sub-directory using the same name:

- **01_Build_Biomass_LUT** – model used to link IPCC Codes with the combined biomass lookup table along with lookup table for growth to produce a master biomass lookup table.
- **02_Load_Landfire_Rasters** – model procedures for loading biomass and carbon values to raw combined (i.e., EVT, EVH, and EVC) LANDFIRE datasets.
- **03_Urban_County_MTCha** – Model procedures used to combine county boundaries with urban carbon values; creates a raster layer that contains mean carbon values for each California county (ARB_URBAN_County_MTCha).
- **04_Calculate_Biomass_MTCha** – model procedures for calculating carbon values by cell and county for the study area. Combines values (total metric tonnes per hectare, including adding in tree growth) for urban, agriculture and wildlands and creates multiple raster files for each LANDFIRE product year in a geodatabase (B04_MAP_MTCha.gdb).
- **05_Calc_Total_Carbon** – model convert carbon densities (MTC/ha) into total carbon values (Metric tonnes of carbon), and summarizes values by different classifications of interest including agriculture, urban, wildland, and IPCC landuse types. Produces per year values.
- **06_Forest_Growth** – model used to calculate carbon values associated with forest growth across years. Adds growth coefficient (6%) to large tree category and generates a raster showing where growth has occurred in the state.
- **07_Net_Carbon** – model that calculates carbon values and rasters associated forest growth to produce a net carbon estimate across years and the study area.

- **08_IPCC_CODE** - model that calculates carbon values and rasters associated landuse conversions across years and the study area. Carbon values (million metric tonnes of carbon) associated with land conversions are captured in a spreadsheet (IPCC_Summary_2010_2001.xlsx).
- **09_Load_Disturbance_Data** – model used to generate disturbance codes and rasters associated with different timber management activities for each year between 2001 and 2010.
- **10_Disturbance_Zone** – combines rasters for each year and identifies the years with a prioritized value associated with timber management. Also combine public and private ownership attributes. Generates a raster attributed with priority timber management activity. This analysis/model needs more investigation to flesh out how best to prioritize management activities.
- **11_Disturbance_Stats** – model used to run zonal statistics to combine all disturbances into geodatabase. Attributes are used to summarize timber management and disturbances (in acres) across years in a spreadsheet (ARB_HARVEST_FIRE_ANNUAL_SUMMARY.xlsx).

The models should be run in sequence where the input data is pulled from the directory of the previous model. This segmentation in the analytical process allows for the user to test individual steps without having to recreate the entire simulation. This approach also allows for all the intermediate products to be saved so they can potentially be used for additional analysis. Also note that the models housed within the ArcGIS geodatabase (i.e., 'Toolbox.tbx') clearly illustrate inputs, procedures and outputs for each model listed immediately above

A oral narrative that contains procedures for using the update GHG Inventory Tool along with a description of all files and folders are also captured in a webinar recording titled “2015-11-19 10.22 ARB Accounting Update.mp4” and is included as a deliverable for this project.

Step 9: Conduct Carbon Stock Change Evaluation

Information and products (e.g., databases, spreadsheets and GIS data) generated through Steps 1 through 8 describe the procedures for organizing data and information for calculating estimates of above ground carbon stock change in California for a given LANDFIRE product year (2001, 2008, and 2010). Using the updated GHG Inventory Tool (in step 8), total above ground live carbon was preliminarily estimated to be about 2,696 MMTC in 2001, and 2551 MMTC in 2010, representing an overall loss of about -145 MMTC over the time period (Table 22) or a loss of approximately -16.1 MMTC yr⁻¹. The greatest estimated loss in carbon pools converting to grasslands with wetlands remaining relatively unchanged across 2001 and 2010 (Table 22). These estimates include above ground live biomass associated with forestlands, croplands, grasslands, wetlands, settlements, and other lands. However, it is important to note that these estimates are preliminary and should not be formally or informally reported. The estimates have not been adjusted for biomass burning, wildfire emissions, or harvested wood products. Forestlands represent the largest carbon pool within the study area, storing about 11 times more carbon than other land categories combined. These adjustments have been incorporated or addressed as separate elements.

Table 22. Preliminary estimates of total above ground live and dead carbon (not including soil carbon) in 2001 and 2010 and associated net carbon change by IPCC land category within California (estimated in MMTC) using the update GHG inventory tool.

Category Name	2001 Above Ground Live Carbon (MMTC)	2010 Above Ground Live Carbon (MMTC)	Net Carbon Stock Change (2001-2010)
Forestland	2,477	2,468	-9
Cropland	41	42	0
Grassland	138	27	-111
Wetlands	0	0	0
Settlement (Urban)	10	7	-3
Other Lands	30	7	-23
Grand Total	2,696	2,551	-145

In addition, calculations for changes through time (2001 to 2010) in net above ground live biomass and carbon values for California as result of land conversions were made using IPCC typology and presented in Table 23. This analysis was conducted after all biomass and carbon values were included in the “ARB_C_v2.7.accdb.” According to the analysis, the largest reduction in net above ground live carbon for this time period across wildland (i.e., forests and other lands), agriculture and urban landscapes was the conversion of the forest type to the grassland type, and the greatest gain in above ground live carbon was the conversion of the wetland type to the forest type (Table 23).

Table 23. Estimated net changes in above ground live and dead biomass associated carbon (MMTC) for land conversions occurring in California from 2001 to 2010 by IPCC categories and subcategories.

Category Name	Sub Category Name	Net Change in Above-Ground Live Biomass Pool (MMTC)	Net Change – All Pools (MMTC)
3B1 Forestland	3B1a Forestland Remaining Forestland	17.50	(16.85)
	3B1bi Cropland Converted to Forestland	0.00	(0.00)
	3B1bii Grassland Converted to Forestland	0.38	3.45
	3B1biii Wetlands Converted to Forestland	0.89	4.19
	3B1biiii Settlements Converted to Forestland	-	-
	3B1bv Other Land Converted to Forest Land	0.01	(0.07)
3B1 Forestland Sub-Total		18.78	(9.27)
3B2 Cropland	3B2a Cropland Remaining Cropland	-	7.99
	3B2bi Forest Converted to Cropland	(1.54)	(7.54)
	3B2bii Grassland Converted to Cropland	(0.11)	(0.19)
	3B2biii Wetlands Converted to Cropland	(0.05)	0.12
	3B2biiii Settlements Converted to Cropland	-	-
	3B2bv Other Land Converted to Cropland	(0.00)	0.02
3B2 Cropland Sub-Total		(1.70)	0.40
3B3 Grassland	3B3a Grassland Remaining Grassland	0.34	1.75
	3B3bi Forest Converted to Grassland	(35.44)	(112.49)
	3B3bii Cropland Converted to Grassland	0.00	(0.00)
	3B3biii Wetlands Converted to Grassland	0.03	0.14
	3B3biiii Settlements Converted to Grassland	-	-
	3B3bv Other Land Converted to Grassland	0.00	0.00
3B3 Grassland Sub-Total		(35.07)	(110.60)
3B4 Wetlands	3B4ai Peatlands Remaining Peatlands	-	-
	3B4aii Flooded Land Remaining Flooded Land	0.00	0.00
	3B4bi Land Converted for Peat Extraction	-	-

Category Name	Sub Category Name	Net Change in Above-Ground Live Biomass Pool (MMTC)	Net Change – All Pools (MMTC)
	3B4bii Land Converted to Flooded Land	-	-
	3B4biii Land Converted to Other Wetland	(0.00)	(0.01)
3B4 Wetlands Sub-Total		(0.00)	(0.01)
3B5 Settlements	3B5a Settlements Remaining Settlements	-	(3.25)
	3B5bi Forestlands Converted to Settlements	(0.11)	(0.52)
	3B5bii Cropland converted to Settlements	-	0.84
	3B5biii Grassland converted to Settlement	(0.01)	(0.02)
	3B5biiii Wetlands converted to Settlement	(0.00)	0.03
	3B5bv Other Land Converted to Settlement	(0.00)	0.00
3B5 Settlements Sub-Total		(0.13)	(2.92)
3B6 Other Land	3B6a Other Land Remaining Other Land	0.01	0.03
	3B6bi Forestland Converted to Other Land	(4.65)	(22.46)
	3B6bii Cropland Converted to Other Land	0.01	(0.07)
	3B6biii Grassland Converted to Other Land	(0.02)	(0.09)
	3B6biiii Wetlands Converted to Other Land	(0.00)	(0.01)
	3B6bv Settlements Converted to Other Land	-	-
3B6 Other Land Sub-Total		(4.66)	(22.61)
3C1 Emissions from biomass Burning	3C1a Biomass Burning in Forestlands	-	-
	3C1b Biomass Burning in Croplands	-	-
	3C1c Biomass Burning in Grasslands	-	-
	3C1d Biomass Burning in Other Lands	-	-
3C1 Emissions from Biomass Burning Total		-	-
3D1 Harvested Wood Products	3D1 Harvested Wood Products	-	-
3D1 Harvested Wood Products Total		-	-
Grand Total		(22.77)	(145.01)

CONCLUSIONS, RECOMMENDATIONS AND NEXT STEPS

Managing Typology Changes in LANDFIRE Existing Vegetation Types. LANDFIRE is an evolving product that is expanding its capacity in resource management beyond wildfires. It is co-funded by two federal agencies (US Department of Agriculture and US Department of the Interior). Thus it has many constituents. In each revision, it tries to respond to requests for a variety of improvements. Also a founding principle in regard to its vegetation mapping was to abide by guidelines in the National Vegetation Classification System. In some respects, LANDFIRE has become the national vegetation map by default. However, as a consequence, continual modification of EVT's is likely as constituent needs and standards change.

For the updated carbon stock assessment tool, our biomass classes are based on the 2008 EVT's. As noted, the trend is to produce more finely resolved vegetation classes and LANDFIRE is committed to a hierarchical approach. The new EVT's will fit under the coarser 2008 classes making it possible to create cross-walks that maintain the consistency of the carbon accounting over time.

The refinements in the 2010 EVT's do suggest the potential for more precise carbon estimation. For the blue oak woodland case study, divisions based on vegetation structure were more relevant for carbon estimation than ones based on species composition. Certainly dividing EVT's that cross physiognomic gradients (e.g., woodland/savanna or woodland/shrubland) into more structurally consistent classes would also reflect gradients in carbon storage. However, we found that even the more cohesive units contained a great deal of plot-to-plot variation in above-ground live biomass. Moreover, results from the vegetation mapping assessment of LANDFIRE (NatureServe 2012) document the challenge in differentiating discrete groups when the vegetation itself is very heterogeneous. In short, LANDFIRE can assign coarser scale vegetation classes with much greater accuracy. Given that the major source of uncertainty in the statewide carbon assessment was LANDFIRE classification (Battles *et al.* 2014), we recommend against recalculating biomass classes for the refined EVT's. In fact, our evaluation suggests that we could gain consistency and reduce uncertainty without a major fall-off in precision by estimating carbon stores as a function of LANDFIRE subclass (e.g., closed-canopy, evergreen forest, sparse canopy mixed forests, open canopy deciduous forest). These more coarse-scale designations are more reliably determined by LANDFIRE and could be segregated by major ecological regions in California to parse major carbon density gradients. For example, the closed-canopy evergreen forest in the north coast supersection would be one group and the closed-canopy evergreen forest in the Sierra Nevada another. Logistically, it would make the most sense to consider this alternative approach the next time new biomass classes are introduced by the LANDFIRE program.

Dead Wood Carbon Pools. LANDFIRE's 100-hr fuels, whether a component of a FBFM or a fuelbed, were most consistently close to field-measured values of all the dead carbon pools. The current comparison only included mixed conifer forest plots in the Sierra Nevada. Field data from other forest types and other regions of California are needed. Additionally, all the datasets are best matched with the 2008 version of LANDFIRE. Several FCCS fuelbeds and FBFMs dominate the LANDFIRE mapping at these sites and have an outsized influence on the comparison, including:

- FCCS
 - 37 Ponderosa pine-Jeffrey pine forest
 - 17 Red fir forest

- 627 Modified or Managed Xeric Understory 2 (based on FBFM TU5 , “very high load, dry climate timber-shrub”)
- 7 Douglas-fir-Sugar pine-Tan oak forest
- FBFM40 values
 - 165/TU5 (“very high load, dry climate timber-shrub”) and 186/TL6 (“moderate load broadleaf litter”)

Above Ground Carbon Stock Changes Analysis (2001 to 2010). The carbon stock change estimates represented in this report are the result of testing different elements (e.g., tables, models) of the updated GHG Inventory Tool. Consequently, the estimates should be considered preliminary and **should not** be represented as an official or qualified accounting of above ground carbon stock change for the state as additional refinement of assumptions and inputs are needed to be informed by ARB staff as appropriate.

Next Steps. The following next steps were identified for additionally refining the GHG Inventory Tool and ARB’s effort to assess above ground carbon stock changes. Specifically, ARB may consider working toward the acquisition of annualized input/base data. Although LANDFIRE has been invaluable for providing good estimates of above ground biomass and associated carbon across the state for natural and working lands, LANDFIRE does not provide annual updates making regular assessment of carbon stocks not possible using this data source. Elements of annual updated vegetation and land cover data should also include improved data on:

- 1) wildfire emission and fuel beds,
- 2) tree mortality and rates of mortality (become an ever present issue associated with drought conditions in California),
- 3) improved annual characterization of urban and agricultural biomass,
- 4) improved estimated of climate change induced type conversions – specifically forest to shrub type conversions and
- 5) more frequent characterization of land cover changes and
- 6) improved forest structure characterization (e.g., tree height).

New data initiatives such as NASA’s Joint Emissivity Database Initiative (JEDI) could aid improved characterization of land cover characteristics on an annualized basis.

REFERENCES

- Allen-Diaz, Barbara; Bartolome, James W.; McClaran, Mitchel P. 1999. California oak savanna. In: Anderson, Roger C.; Fralish, James S.; Baskin, Jerry M., eds. Savannas, barrens, and rock outcrop plant communities of North America. New York: Cambridge University Press: 322-339.
- Bjorkman, J., J.H. Thorne, A. Hollander, N.E. Roth, R.M. Boynton, J. de Goede, Q. Xiao, K. Beardsley, G. McPherson, J.F. Quinn. March, 2015. Biomass, carbon sequestration and avoided emission: assessing the role of urban trees in California. Information Center for the Environment, University of California, Davis.
- Boryan, C., Yang, Z., Mueller, R., Craig, M., 2011. Monitoring US agriculture: the US Department of Agriculture, National Agricultural Statistics Service, Cropland Data Layer Program. Geocarto International, 26(5), 341–358. <http://nassgeodata.gmu.edu/CropScape/>
- California (CALFIRE) 2010. Chapter 1.2: Sustainable Working Forests and Rangelands. In: California’s Forests and Rangelands: 2015 Assessment. CALFIRE’s Fire and Resource Assessment Program (FRAP), Sacramento, CA, 34p.
- California State Board of Equalization (BOE) 2015. California Timber Harvest Statistics. <https://www.boe.ca.gov/proptaxes/pdf/harvy2.pdf> Last accessed: November 16 2015.

- Carlisle, E., Smart, D., Williams, L.E., & Summers, M., 2010. California vineyard greenhouse gas emissions: Assessment of the available literature and determination of research needs. California Sustainable Wine Growing Alliance.
- Ciais, P., C. Sabine, G. Bala, L. Bopp, V. Brovkin, J. Canadell, A. Chhabra, R. DeFries, J. Galloway, M. Heimann, C. Jones, C. Le Quéré, R.B. Myneni, S. Piao and P. Thornton, 2013: Carbon and Other Biogeochemical Cycles. In: Climate Change 2013: The Physical Science Basis. Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change [Stocker, T.F., D. Qin, G.-K. Plattner, M. Tignor, S.K. Allen, J. Boschung, A. Nauels, Y. Xia, V. Bex and P.M. Midgley (eds.)]. Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA.
- Collins, B., S. Stephens, G. Roller, and J. Battles. 2011. Simulating fire and forest dynamics for a landscape fuel treatment project in the Sierra Nevada. *Forest Science* 57(2): 77-88.
- DeJong, T. M., 2013. Developing a Carbon Budget, Physiology, Growth and Yield Model for Almond Trees. Almond Board of California. Final Report.
- DOE 2007. Appendices to the Technical Guidelines: Part I Appendix: Forestry. Department of Energy, Washington DC, 280p. <http://www.eia.doe.gov/oiaf/1605/Forestryappendix%5B1%5D.pdf>. Last viewed on September 13 2012
- FRAP 2010. Chapter 1.2 Easter, Mark. 2015. Natural Resource Ecology Laboratory, Colorado State University. Personal communication.
- Eve, M., D. Pape, M. Flugge, R. Steele, D. Man, M. Riley-Gilbert, and S. Biggar, (Eds), 2014. Quantifying Greenhouse Gas Fluxes in Agriculture and Forestry: Methods for Entity-Scale Inventory. Technical Bulletin Number 1939. Office of the Chief Economist, U.S. Department of Agriculture, Washington, DC. July 2014.
- FGDC. 2008. Federal Geographic Data Committee. National Vegetation Classification System. Version 2. FGDC-STD-005-2008 (Version 2).
- Forest Inventory Analysis (FIA) 2015. EVALIDator Version 1.6.0.03. McCune, B., J.B. Grace, and D.L. Urban. 2002. [Analysis of Ecological Communities](#). Glenden Beach, OR: MjM Software Design.
- IPCC 2013
- McNab, W.H.; Cleland, D.T.; Freeouf, J.A.; Keys, Jr., J.E.; Nowacki, G.J.; Carpenter, C.A., comps. 2005. Description of ecological subregions: sections of the conterminous United States [CD-ROM]. Washington, DC: U.S. Department of Agriculture, Forest Service. 80 p.
- NatureServe. 2012. Expert Attribution for Auto-Key Improvements (LANDFIRE) and Advancing Methods for integration with the revised US-National Vegetation Classification Standard: GeoArea 3. Final Report: Boulder, CO.
- NVCS. 2015. National Vegetation Classification Standard. <https://www.fgdc.gov/standards/projects/FGDC-standards-projects/vegetation>. <http://apps.fs.fed.us/Evalidator/evalidator.jsp> Last accessed: November 16 2015.
- LANDFIRE 2015. Disturbance 1999-2012, Fuel Disturbance, Vegetation Disturbance. http://www.landfire.gov/disturbance_2.php Last accessed: November 16 2015.
- LANDFIRE 2016. LANDFIRE Data Dictionary. http://www.landfire.gov/DataDictionary/lf_vdist2012.xml Last accessed: January 3 2016.
- Li, S., Potter, C., & Hiatt, C., 2012. Monitoring of net primary production in California rangelands using Landsat and MODIS satellite remote sensing. *Scientific Research, Natural Resources*, Vol. 3, 56-65. DOI: 10.4236/nr.2012.32009.
- McGroddy, M.E., T. Daufresne, and L.O. Hedin. 2004. Scaling of C:N:P stoichiometry in forests worldwide: Implications of terrestrial Redfield-type ratios. *Ecology* 85: 2390-2401.
- McIver, CP, Meek, JP, Scudder, MG, Sorenson, CB, Morgan, TA, Christensen, GA, 2015. DRAFT - In Production California's Forest Products Industry and Timber Harvest, 2012. PNW-GTR-908. USDA Forest Service, Pacific Northwest Research Station, 53p.

- Mitchell, J., Hartz, T., Pettygrove, S., Munk, D., May, D., Menezes, F., Diener, J., & O'Neill, T. 1999. Organic matter recycling varies with crops grown. *California Agriculture*, 53(4), 37-40.
- Morgan, K. T., Scholberg, J. M. S., Obreza, T. A., & Wheaton, T. A. 2006. Size, biomass, and nitrogen relationships with sweet orange tree growth. *Journal of the American Society for Horticultural Science*, 131(1), 149-156.
- Murphy, William J., 1994. University of Missouri Extension, Department of Agronomy, G4020, Tables for Weights and Measurements: Crops. Publication: #G4020. <http://extension.missouri.edu/publications/DisplayPub.aspx?P=G4020>
- National Agricultural Statistics Service. <http://quickstats.nass.usda.gov/>
- Ottmar, R.D.; Sandberg, D.V.; Riccardi, C.L.; Prichard, S.J. 2007. An overview of the fuel characteristic classification system – quantifying, classifying, and creating fuelbeds for resource planning. *Canadian Journal of Forest Research*. 37(12): 2383-2393.
- Penman, J., et. al 2003. Intergovernmental Panel on Climate Change (IPCC). Good practice guidance for land use, land-use change and forestry. Institute for Global Environmental Strategies. http://www.ipcc-nggip.iges.or.jp/public/gpplulucf/gpplulucf_contents.html.
- Prichard, S. J., Sandberg, D. V., Ottmar, R. D., Eberhardt, E., Andreu, A., Eagle, P., Swedin, K.. 2013. Fuel Characteristic Classification System version 3.0: technical documentation. Gen. Tech. Rep. PNW-GTR-887. Portland, OR: U.S. Department of Agriculture, Forest Service, Pacific Northwest Research Station. 79 p. http://www.fs.fed.us/pnw/pubs/pnw_gtr887.pdf.
- Putnam, Daniel H. 2015. Alfalfa and Forage News. UC Cooperative Extension: <http://ucanr.edu/blogs/blogcore/postdetail.cfm?postnum=17721>
- Rosecrance, R. and Lovatt, C. 2003. Seasonal Patterns of Nutrient Uptake and Partitioning as a Function of Crop Load of the 'Hass' avocado. CSU, Chico. Final Report.
- Ruesch, Aaron, and Holly K. Gibbs. 2008. New IPCC Tier-1 Global Biomass Carbon Map for the Year 2000. Carbon Dioxide Information Analysis Center [<http://cdiac.ornl.gov>], Oak Ridge National Laboratory, Oak Ridge, Tennessee. http://cdiac.ornl.gov/epubs/ndp/global_carbon/carbon_documentation.html
- Ryals, R., & Silver, W. L. 2013. Effects of organic matter amendments on net primary productivity and greenhouse gas emissions in annual grasslands. *Ecological Applications*, 23(1), 46-59.
- Saah, D, Robards, T, Moody, T., O'Neil-Dune, J., Moritz, M., Hurteau, M., Moghaddas, J. 2012. Developing an Analytical Framework for Quantifying Greenhouse Gas Emission Reductions from Forest Fuel Treatment Projects in Placer County, California. Prepared for: United States Forest Service: Pacific Southwest Research Station, 130p.
- Skog, KE, Nicholson, GA 2000. Carbon sequestration in wood and paper products. In: Joyce, Linda A.; Birdsey, Richard, technical editors. 2000. The impact of climate change on America's forests: a technical document supporting the 2000 USDA Forest Service RPA Assessment. Gen. Tech. Rep. RMRS-GTR-59. Fort Collins, CO: U.S. Department of Agriculture, Forest Service, Rocky Mountain Research Station. p. 79-88
- Smith, J.E., L.S. Heath, K.E. Skog, and R.A. Birdsey. 2006. Methods for calculating forest ecosystem and harvested carbon with standard estimates for forest types of the United States. Newtown Square, PA: US Department of Agriculture, Forest Service, Northern Research Station.
- University of California 2015. Carbon Sequestration Tool for THPs. http://ucanr.edu/sites/forestry/Carbon_Sequestration_Tool_for_THPs/
- USFS. 2013. Baseline assessment of forest carbon stocks including harvested wood products – USDA Forest Service, Pacific Southwest Region. Report from the Climate Change Advisor's Office, Office of the Chief.

- West, T. O., Brandt, C. C., Baskaran, L. M., Hellwinckel, C. M., Mueller, R., Bernacchi, C. J., Bandaru, V., Yang, B., Wilson, B. S., Marland, G., Nelson, R. G., De La Torre Ugarte, D. G., & Post, W. M. 2010. Cropland carbon fluxes in the United States: increasing geospatial resolution of inventory-based carbon accounting. *Ecological Applications*, 20(4), 1074-1086.
- Zheng, D. L. S. Heath, M.J. Ducey and J. E. Smith. 2011. Carbon changes in conterminous US forests associated with growth and major disturbances: 1992–2001. *Environmental Resource Letters* 6: 014012.

APPENDICES

Appendix 1. List and summary description of documents, spreadsheets, lookup tables, databases, and geodatabases used to complete supporting analysis and update of GHG Inventory Tool.

File Name	Application In ARB Project	Description
ca_105_FCCS_FBFM40_FIA_output.xls	Task 1 - dead wood analysis	The “comparison” tab of this spreadsheet shows the average difference (in tC/ac) between current FIA dead carbon pools and 2001 LANDFIRE Scott & Burgan fuel model dead carbon pools and 2001 LANDFIRE FCCS fuelbed dead carbon pools for 23 ecological subregions.
ca_110_FCCS_FBFM40_FIA_output.xls	Task 1 - dead wood analysis	The “comparison” tab of this spreadsheet shows the average difference (in tC/ac) between current FIA dead carbon pools and 2008 LANDFIRE Scott & Burgan fuel model dead carbon pools and 2008 LANDFIRE FCCS fuelbed dead carbon pools for 23 ecological subregions.
ARB GHG Task 1 raw data and pivot.xls	Task 1 - dead wood analysis	The “all raw data” tab contains the dead carbon pool data for about 2,000 field plots. The “pivot” tab formats this data into a pivot table. The “stats” tab summarizes the difference between 2008 LANDFIRE Scott & Burgan fuel model dead carbon pools and field plot dead carbon pools as well as 2008 LANDFIRE Scott & Burgan FCCS fuelbed dead carbon pools and field plot dead carbon pools.
FIA comparison documentation.docx	Task 1 - dead wood analysis	Provides a description of steps used to conduct dead wood analysis
ARB GHG task 1 preliminary report.docx	Task 1 – dead wood analysis	Interim report presented to ARB on July 31, 2015 that describes methods and results of comparison of FCCS and Scott and Burgan Fuels models to forest plot data.
Ag and Urban Carbon estimates by EVT.xls	Task 2 – agriculture and urban carbon	Above ground carbon estimates and IPCC defaults by LandFire EVT’s.
Ag and Urban Review 2010 EVTs_20150916.xls	Task 2 – agriculture and urban carbon	Pixel count and areas of CA specific LandFire EVT’s. Crop specific carbon calculations and references.
Field crop estimates_20150916.xls	Task 2 – agriculture and urban carbon	Complete listing of carbon estimates by individual crop.
Urban biomass_CA county_20150916.xls	Task 2 – agriculture and urban carbon	Urban biomass and carbon estimates by CA county.

File Name	Application In ARB Project	Description
ARB Task 2 Lit review-Master_20150917.accdb	Task 2 – agriculture and urban carbon	Contains citations, sources, and biomass and carbon values from urban and agriculture literature review.
Ag and Urban Carbon estimates by EVT_v5_10062015.xls	Task 2 – agriculture and urban carbon	Summary tables (in different tabs) for biomass and carbon estimates (above and below ground) derived from literature review for county, crop type and EVT.
Task2_ag_and_other_lands_Carbon_dbasetable_10122015.xls	Task 2 – agriculture and urban carbon	Above ground carbon values (MgC/ha) for agriculture and other lands summarized by county and EVT
EVT_Cdata_xwalk_10242015v3.xls	Task 2 – agriculture and urban carbon	Peak above ground carbon by EVT, year, and county for urban and agriculture landscapes.
ARB_ForestSectorGHG_Enhancement_14-757_Prog_Report_Oct_12_2015_Gunn Revised.ppt	Task 2 – agriculture and urban carbon	Powerpoint presentation presented at October 12, 2015 meeting with ARB – summarizes status of dead wood analysis (Task 1), agriculture and urban literature review (Task 2), and timber management LCA and associated analysis (Task 4).
PC173-ARB Task 2 Interim Report_September 17_2015_Draft.docx	Task 2 – agriculture and urban carbon	Interim report for Task 2 (urban and agriculture carbon literature review) presented at September 18 th Project team meeting.
2010 Battles XWALK details.xls	Task 3 – Biomass Classes Lookup Table Update	A crosswalk of LANDFIRE existing vegetation types (EVT) across 2001, 2008 and 2010).
ARB_C_LUT_v2.7.accdb	Task 3 – Biomass Classes Lookup Table Update	Access database that is the foundation of the updated GHG Inventory Tool. Contains lookup tables, and associated biomass and carbon estimates for natural, urban and agricultural landscapes for 2001, 2008 and 2010, as well as for land conversions.
CrossWalk 2001 2008 2010 EVT to Biomass Class.xls	Task 3 – Biomass Classes Lookup Table Update	Crosswalk of LANDFIRE EVT category and IPCC landuse category by year (2001, 2008, 2010) including notes for certain vegetation types.

File Name	Application In ARB Project	Description
BATTLES_BIOMASS_LUT_ALL Revised.xls	Task 3 – Biomass Classes Lookup Table Update	Final updated biomass class lookup table. Contains tab that summarizes of stock changes between 2001, 2008 and 2010 (includes stock changes associates with urban and agriculture landscapes).
ARB_IPCC_TABLE_20151102.xls	Task 3 – Biomass Classes Lookup Table Update	Spreadsheet that summarizes above-ground live carbon (MMTC) stock changes associated with land conversion using IPCC landuse categories.
Battles Progress Report Carbon Cross Walk (003).ppt	Task 3 – Biomass Classes Lookup Table Update	Powerpoint presentation on analysis completed to understand how changes in LANDFIRE EVT's might affect carbon stock change assessment. Presented at September 18, 2015 Project Team meeting by Dr. John Battles.
LUT_ARB_BIOMASS_v2_7.xls	Task 3 – Biomass Classes Lookup Table Update	Has all EVT biomass values including urban and agriculture. Per pixel carbon values. Including IPCC categories. A macro in ARB_C_LUT_v2.7.accdb generated this spreadsheet.
LUT_GUNN_Urban_CNTYv4.xls	Task 3 – Biomass Classes Lookup Table Update	Contains MTC/hectare values by county and source of values. A macro in ARB_C_LUT_v2.7.accdb generated this spreadsheet.
LUT_IPCC_Activity.xls	Task 3 – Biomass Classes Lookup Table Update	Contains IPCC landuse categories and combinations of land conversions (start landuse to end landuse codes). A macro in ARB_C_LUT_v2.7.accdb generated this spreadsheet. There is also a corresponding table in the .gdb.
LUT_IPCC_CODEv1.xls	Task 3 – Biomass Classes Lookup Table Update	A generic lookup table that provides codes for different IPCC landuse categories. A macro in ARB_C_LUT_v2.7.accdb generated this spreadsheet.

File Name	Application In ARB Project	Description
2015-11-19 10.22 ARB Accounting Update.mp4	Task 3 – Biomass Classes Lookup Table Update	Recording of SIG GoTo Meeting with ARB on November 19, 2015. In the recording, David Saah describes and systematically walks through the different databases, geographic data (geodatabases), scripts (macros) and lookup tables used in the updated GHG inventory tool.
LUT_LUT_Disturbance.xls	Task 4 – LCA for Forest Management	Spreadsheet mostly related to Tasks 4. Contains database lookup codes for different types of disturbances (from the LANDFIRE disturbance layer). A macro in ARB_C_LUT_v2.7.accdb generated this spreadsheet.
ARB TASK 4 methods and result writeup 2015-11-16.docx	Task 4 – LCA for Forest Management	Interim report on methods and results related to harvest associated carbon losses on the landscape across California from 2001 to 2010.
Disturbance_2010_2001 2015-11-16.xls	Task 4 – LCA for Forest Management	Summary of harvest/disturbance area (by type) and associated biomass and carbon by county and ownership - derived from LANDFIRE Disturbance layer (1999 to 2012). This spreadsheet contains a pivot table tool that provides options for querying carbon values for different disturbance types, including: 1) development, 2) timber harvest, 3) insects, 4) prescribed fire, 5) wildland fire, 6) disease, 7) herbicide, 8) mastication, 9) other mechanical, and 10) wildfire use.
Wood product C pools from CA 1999-2012 2015-10-16.xls	Task 4 – LCA for Forest Management	Spreadsheet contains a comprehensive summary of natural and anthropogenic disturbances in California from 1999 to 2012 and the fate of biomass and carbon (in-use and post use). Contains a dashboard that summarizes above ground carbon in comparison to 1605(b) and Stewart and Nakamura. Tabs are included for annualized 100 year total harvest wood products (in mg C), Accumulated fate of wood products calculations (in use, in landfills, emissions), Stewart and Nakamura data, above ground carbon in live trees from FIA EVALIDator v1.6.0.03 (0 to 500+ years), wood volume per hectare, CA wood products consumption (1970 to 2010) data, data on the whereabouts of wood products (1970 to 2012), BOE data on harvest volumes (1978 to 2014), Forest Types by ownership and associated volumes (source FIA), USFS PNW Ca. timber harvest (mmbf) by ownership (1952-2008), log exports (mmbf) 1961 to 2012, decadal CA census data (1850 to 2010), timber consumption data (1965 to 2002), Smith et al. 2006 wood product fractional C emissions (96 years), Stewart and Nakamura wood product fate (160 year), ARB CO2 emissions from discarded wood and paper in landfills, CA origin wood products delivered to landfill, Landfilled wood products (tons C) 1970 to 2010, and Acres of standard silvicultural prescriptions on private timberlands in Timber Harvesting Plans by year

File Name	Application In ARB Project	Description
ARB_IPCC_TABLE_20151102.xls	Task 5 – IPCC crosswalk with LANDFIRE EVT	Peak above ground live carbon by EVT, year, and county for urban and agriculture landscapes. Spreadsheet includes estimates of changes in ABL associated with land conversions in California between 2001 and 2010.
LUT_Order_Growthv1.xls	Not in Original Agreement Scope	A database lookup table code for different LANDFIRE Existing Vegetation “Orders”. Used to address undetected growth. A macro in ARB_C_LUT_v2.7.accdb generated this spreadsheet.

Appendix 2. List of sources reviewed for deriving biomass and carbon values for agriculture and urban landscapes.

- Aertsens, J., De Nocker, L., & Gobin, A., 2013. Valuing the carbon sequestration potential for European agriculture. *Land Use Policy*, 31, 584–594. doi:10.1016/j.landusepol.2012.09.003
- Aguilera, E., Lassaletta, L., Gattinger, A., & Gimeno, B. S., 2013. Managing soil carbon for climate change mitigation and adaptation in Mediterranean cropping systems: A meta-analysis. *Agriculture, Ecosystems and Environment*, 168, 25–36. doi:10.1016/j.agee.2013.02.003
- American Carbon Registry, 2013. Voluntary Emission Reductions in Rice Management Systems, v1.0. Terra Global Capital, LLC.
- Bajocco, S., Dragoz, E., Gitas, I., Smiraglia, D., Salvati L., Ricotta, C., 2015. Mapping Forest Fuels through Vegetation Phenology: The Role of Coarse-Resolution Satellite Time-Series. *PLoS ONE* 10(3): e0119811. doi:10.1371/journal.pone.0119811.
- Bjorkman, J., J.H. Thorne, A. Hollander, N.E. Roth, R.M. Boynton, J. de Goede, Q. Xiao, K. Beardsley, G. McPherson, J.F. Quinn. March, 2015. Biomass, carbon sequestration and avoided emission: assessing the role of urban trees in California. Information Center for the Environment, University of California, Davis.
- Blackard, J. A., Finco, M. V., Helmer, E. H., Holden, G. R., Hoppus, M. L., Jacobs, D. M., Lister, A. J., Moisen, G. G., Nelson, M. D., Riemann, R., Ruefenacht, B., Salajanu, D., Weyermann, D. L., Winterberger, K. C., Brandeis, T. J., Czaplewski, R. L., McRoberts, R. E., Patterson, P. L., & Tymcio, R. P., 2008. Mapping US forest biomass using nationwide forest inventory data and moderate resolution information. *Remote Sensing of Environment*, 112(4), 1658-1677.
- Brown, S.T., A. Pearson, J. Dushku, J. Kadyzewski, and Qi, Y., 2004. Baseline greenhouse gas emissions and removals of forest, range, and agricultural lands in California. Winrock International, for the California Energy Commission, PIER report 500-04-069F. 80p.
- Buchholz, T., Hurteau, M. D., Gunn, J., & Saah, D., 2015. A global meta-analysis of forest bioenergy greenhouse gas emission accounting studies. *GCB Bioenergy*. Doi: 10.1111/gcbb.12245.
- Byrd, K. B., Flint, L. E., Alvarez, P., Casey, C. F., Sleeter, B. M., Soulard, C. E., Flint, A. L., & Sohl, T. L., 2015. Integrated climate and land use change scenarios for California rangeland ecosystem services: wildlife habitat, soil carbon, and water supply. *Landscape Ecology*, 30(4), 729-750.
- California Agricultural Statistics Review, 2013-2014. California Department of Food and Agriculture. Sacramento, CA.
- Carlisle, E., Smart, D., Williams, L.E., & Summers, M., 2010. California vineyard greenhouse gas emissions: Assessment of the available literature and determination of research needs. California Sustainable Wine Growing Alliance.
- Clark, D. A., Brown, S., Kicklighter, D. W., Chambers, J. Q., Thomlinson, J. R., & Ni, J., 2001. Measuring net primary production in forests: concepts and field methods. *Ecological applications*, 11(2), 356-370.
- Cleveland, C. C., Taylor, P., Chadwick, K. D., Dahlin, K., Doughty, C. E., Malhi, Y., Smith, W. K., Sullivan, B. W., Wieder, W. R., & Townsend, A. R. 2015. A comparison of plot-based, satellite and Earth system model estimates of tropical forest net primary production. *Global Biogeochemical Cycles*. DOI: 10.1002/2014GB005022
- Conant, R. T., Paustian, K., & Elliott, E. T., 2001. Grassland management and conversion into grassland: Effects on soil carbon. *Ecological Applications*, 11(2), 343–355.
- DeFries, R., Achard, F., Brown, S., Herold, M., Murdiyarso, D., Schlamadinger, B., & de Souza Jr., C., 2007. Earth observations for estimating greenhouse gas emissions from deforestation in developing countries. *Environmental science & policy*, 10(4), 385-394.
- DeJong, T. M., 2013. Developing a Carbon Budget, Physiology, Growth and Yield Model for Almond Trees. Almond Board of California. Final Report.
- Domke, G. M., Woodall, C. W., & Smith, J. E., 2011. Accounting for density reduction and structural loss in standing dead trees: Implications for forest biomass and carbon stock estimates in the United States. *Carbon balance and management*, 6(1), 1-11.

- Dore, S., Montes-Helu, M., Hart, S. C., Hungate, B. A., Koch, G. W., Moon, J. B., Finkral, A. G., & Kolb, T. E., 2012. Recovery of ponderosa pine ecosystem carbon and water fluxes from thinning and stand-replacing fire. *Global Change Biology*, 18(10), 3171-3185.
- Eve, M., D. Pape, M. Flugge, R. Steele, D. Man, M. Riley-Gilbert, and S. Biggar, (Eds), 2014. Quantifying Greenhouse Gas Fluxes in Agriculture and Forestry: Methods for Entity-Scale Inventory. Technical Bulletin Number 1939. Office of the Chief Economist, U.S. Department of Agriculture, Washington, DC. July 2014.
- Finkral, A. J., & Evans, A. M., 2008. The effects of a thinning treatment on carbon stocks in a northern Arizona ponderosa pine forest. *Forest Ecology and Management*, 255(7), 2743-2750.
- Friedl, M. a., Sulla-Menashe, D., Tan, B., Schneider, A., Ramankutty, N., Sibley, A., & Huang, X., 2010. MODIS Collection 5 global land cover: Algorithm refinements and characterization of new datasets. *Remote Sensing of Environment*, 114(1), 168–182. doi:10.1016/j.rse.2009.08.016
- Gibbs, H., Yui, S., & Plevin, R. 2014. New estimates of soil and biomass carbon stocks for global economic models (No. 4344). Center for Global Trade Analysis, Department of Agricultural Economics, Purdue University, Technical paper #33.
- Gibbs, H., Yui, S., & Plevin, R., 2014. New Estimates of Soil and Biomass Carbon Stocks for Global Economic Models (No. 4344). Center for Global Trade Analysis, Department of Agricultural Economics, Purdue University.
- Goetz, Scott J., Baccini, A., Laporte, N.T., Johns, T., Walker, W., Kellndorfer, J., Houghton, R. A., & Sun, M., 2009. Mapping and monitoring carbon stocks with satellite observations: a comparison of methods. *Carbon balance and management* 4.1
- Gonzalez, P., Asner, G. P., Battles, J. J., Lefsky, M. A., Waring, K. M., & Palace, M., 2010. Forest carbon densities and uncertainties from Lidar, QuickBird, and field measurements in California. *Remote Sensing of Environment*, 114(7), 1561-1575.
- Guo, L. B., & Gifford, R. M., 2002. Soil carbon stocks and land use change: a meta analysis. *Global change biology*, 8(4), 345-360.
- Heath, L. S., Smith, J. E., & Birdsey, R. A., 2003. Carbon trends in US forest lands: a context for the role of soils in forest carbon sequestration (pp. 35-45). CRC Press, New York.
- Hicke, J. A., Meddens, A. J., Allen, C. D., & Kolden, C. A., 2013. Carbon stocks of trees killed by bark beetles and wildfire in the western United States. *Environmental Research Letters*, 8(3), 035032.
- Houghton, R. A., Hackler, J. L., Lawrence, K. T., 1999. The U.S. Carbon Budget: Contributions from Land-Use Change. *Science*, 285, 574–578. doi:10.1126/science.285.5427.574.
- Huang, C., Goward, S. N., Masek, J. G., Thomas, N., Zhu, Z., & Vogelmann, J. E., 2010. An automated approach for reconstructing recent forest disturbance history using dense Landsat time series stacks. *Remote Sensing of Environment*, 114(1), 183-198.
- Hudiburg, T., Law, B., Turner, D. P., Campbell, J., Donato, D., & Duane, M., 2009. Carbon dynamics of Oregon and Northern California forests and potential land-based carbon storage. *Ecological applications*, 19(1), 163-180.
- Hurteau, M., & North, M., 2008. Fuel treatment effects on tree-based forest carbon storage and emissions under modeled wildfire scenarios. *Frontiers in Ecology and the Environment*, 7(8), 409-414.
- Jin, S., Yang, L., Danielson, P., Homer, C., Fry, J., & Xian, G., 2013. A comprehensive change detection method for updating the National Land Cover Database to circa 2011. *Remote Sensing of Environment*, 132, 159–175. doi:10.1016/j.rse.2013.01.012
- Kim, D. G., & Kirschbaum, M. U., 2015. The effect of land-use change on the net exchange rates of greenhouse gases: A compilation of estimates. *Agriculture, Ecosystems & Environment*, 208, 114-126.
- Koteen, L. E., Baldocchi, D. D., & Harte, J., 2011. Invasion of non-native grasses causes a drop in soil carbon storage in California grasslands. *Environmental Research Letters*, 6(4), 044001.
- Kroodsma, D. A., & Field, C. B. 2006. Carbon sequestration in california agriculture, 1980-2000. *Ecological Applications*, 16(5), 1975-1985.

- Kurz, W. A., Dymond, C. C., White, T. M., Stinson, G., Shaw, C. H., Rampley, G. J., Smyth, C., Simpson, B. N., Neilson, E. T., Trofymow, J. A., Metsaranta, J., & Apps, M. J., 2009. CBM-CFS3: A model of carbon-dynamics in forestry and land-use change implementing IPCC standards. *Ecological Modelling*, 220, 480–504. doi:10.1016/j.ecolmodel.2008.10.018.
- Lal, R., 2005. Forest soils and carbon sequestration. *Forest Ecology and Management*, 220, 242–258. doi:10.1016/j.foreco.2005.08.015
- Lark, T. J., Salmon, J. M., & Gibbs, H. K., 2015. Cropland expansion outpaces agricultural and biofuel policies in the United States. *Environmental Research Letters*, 10(4), 044003.
- Leifeld, J., & Kögel-Knabner, I., 2005. Soil organic matter fractions as early indicators for carbon stock changes under different land-use?. *Geoderma*, 124(1), 143-155.
- Li, S., Potter, C., & Hiatt, C., 2012. Monitoring of net primary production in California rangelands using Landsat and MODIS satellite remote sensing. *Scientific Research, Natural Resources*, Vol. 3, 56-65. DOI: 10.4236/nr.2012.32009.
- Liao, C., Luo, Y., Fang, C., & Li, B., 2010. Ecosystem carbon stock influenced by plantation practice: implications for planting forests as a measure of climate change mitigation. *PloS one*, 5(5), e10867.
- Liska, A. J., Yang, H., Milner, M., Goddard, S., Blanco-Canqui, H., Pelton, M. P., Fang X. X., Zhu, H., & Suyker, A. E. 2014. Biofuels from crop residue can reduce soil carbon and increase CO₂ emissions. *Nature Climate Change*, 4(5), 398-401.
- Liu, J., Vogelmann, J. E., Zhu, Z., Key, C. H., Sleeter, B. M., Price, D. T., Chen, J. M., Cochrane, M. A., Eidenshink, J. C., Howard, S. M., Bliss, N. B., & Jiang, H. 2011. Estimating California ecosystem carbon change using process model and land cover disturbance data: 1951–2000. *Ecological Modelling*, 222(14), 2333-2341.
- Loboda, T., O'neal, K. J., & Csiszar, I., 2007. Regionally adaptable dNBR-based algorithm for burned area mapping from MODIS data. *Remote Sensing of Environment*, 109(4), 429-442.
- Mallinis, G., Galidaki, G., & Gitas, I., 2014. A comparative analysis of EO-1 Hyperion, Quickbird and Landsat TM imagery for fuel type mapping of a typical Mediterranean landscape. *Remote Sensing*, 6(2), 1684-1704.
- Marvinney, E., Kendall, A., Brodt, S., 2014. A comparative assessment of greenhouse gas emissions in California almond, pistachio, and walnut production. *LCA Food. Proceedings of the 9th International Conference on Life Cycle Assessment in the Agri-Food Sector*
- Meng, R., Dennison, P. E., D'Antonio, C. M., & Moritz, M. A. (2014). Remote Sensing Analysis of Vegetation Recovery following Short-Interval Fires in Southern California Shrublands. *PLoS ONE*, 9(10), e110637.
- Mitchell, J., Hartz, T., Pettygrove, S., Munk, D., May, D., Menezes, F., Diener, J., & O'Neill, T. 1999. Organic matter recycling varies with crops grown. *California Agriculture*, 53(4), 37-40.
- Monleon VJ, Lintz HE (2015) Evidence of Tree Species' Range Shifts in a Complex Landscape. *PLoS ONE* 10(1): e0118069. doi:10.1371/journal.pone.0118069
- Morgan, K. T., Scholberg, J. M. S., Obreza, T. A., & Wheaton, T. A. 2006. Size, biomass, and nitrogen relationships with sweet orange tree growth. *Journal of the American Society for Horticultural Science*, 131(1), 149-156.
- Murty, D., Kirschbaum, M. U. F., Mcmurtrie, R. E. and Mcgilvray, H., 2002, Does conversion of forest to agricultural land change soil carbon and nitrogen? A review of the literature. *Global Change Biology*, 8: 105–123. doi: 10.1046/j.1354-1013.2001.00459.x
- North, M., Hurteau, M., & Innes, J., 2009. Fire suppression and fuels treatment effects on mixed-conifer carbon stocks and emissions. *Ecological Applications*, 19(6), 1385-1396.
- Ogle, S. M., Breidt, F. J., Easter, M., Williams, S., & Paustian, K., 2007. An empirically based approach for estimating uncertainty associated with modelling carbon sequestration in soils. *Ecological Modelling*, 205(3), 453-463.
- Pettorelli, N., Vik, J. O., Mysterud, A., Gaillard, J. M., Tucker, C. J., & Stenseth, N. C., 2005. Using the satellite-derived NDVI to assess ecological responses to environmental change. *Trends in Ecology & Evolution*, 20(9), 503-510.

- Plevin, R. J., O'Hare, M., Jones, A. D., Torn, M. S., & Gibbs, H. K., 2010. Greenhouse gas emissions from biofuels' indirect land use change are uncertain but may be much greater than previously estimated. *Environmental Science and Technology*, 44(21), 8015–8021. doi:10.1021/es101946t
- Potter, C., 2012. Net primary production and carbon cycling in coast redwood forests of central California. *Open Journal of Ecology*.
- Putnam, Daniel H. 2015. Alfalfa and Forage News. UC Cooperative Extension: <http://ucanr.edu/blogs/blogcore/postdetail.cfm?postnum=17721>
- Rosecrance, R. and Lovatt, C. 2003. Seasonal Patterns of Nutrient Uptake and Partitioning as a Function of Crop Load of the 'Hass' avocado. CSU, Chico. Final Report.
- Roy, D.P., Ju, J., Kline, K., Scaramuzza, P.L., Kovalsky, V., Hansen, M.C., Loveland, T.R., Vermote, E.F., Zhang, C., 2010, Web-enabled Landsat Data (WELD): Landsat ETM+ Composited Mosaics of the Conterminous United States. *Remote Sensing of Environment* 114: 35-49.
- Ryals, R., & Silver, W. L. 2013. Effects of organic matter amendments on net primary productivity and greenhouse gas emissions in annual grasslands. *Ecological Applications*, 23(1), 46-59.
- Ryan, K. C., & Opperman, T. S., 2013. LANDFIRE - A national vegetation/fuels data base for use in fuels treatment, restoration, and suppression planning. *Forest Ecology and Management*, 294, 208–216. doi:10.1016/j.foreco.2012.11.003
- Seto, K. C., Fragkias, M., Güneralp, B., & Reilly, M. K., 2011. A meta-analysis of global urban land expansion. *PloS ONE*, 6(8), e23777. doi:10.1371/journal.pone.0023777.
- Silver, W. L., Ryals, R., & Eviner, V., 2010. Soil carbon pools in California's annual grassland ecosystems. *Rangeland Ecology & Management*, 63(1), 128-136.
- Smith, J. E., Heath, L. S., & Woodbury, P. B., 2004. How to estimate forest carbon for large areas from inventory data. *Journal of Forestry*, 102(5), 25-31.
- Smith, James E., Heath, Linda S., Skog, Kenneth E., Birdsey, Richard A. 2006. Methods for calculating forest ecosystem and harvested carbon with standard estimates for forest types of the United States. Gen. Tech. Rep. NE-343. Newtown Square, PA: U.S. Department of Agriculture, Forest Service, Northeastern Research Station. 216 p.
- Springsteen, B., Christofk, T., Eubanks, S., Mason, T., Clavin, C., & Storey, B., 2011. Emission reductions from woody biomass waste for energy as an alternative to open burning. *Journal of the Air & Waste Management Association*, 61(1), 63-68. DOI: 10.3155/1047-3289.61.1.63.
- Stevens, A., van Wesemael, B., Vandenschrick, G., Touré, S., & Tychon, B., 2006. Detection of Carbon Stock Change in Agricultural Soils Using Spectroscopic Techniques. *Soil Science Society of America Journal*, 70, 844. doi:10.2136/sssaj2005.0025
- T. M. Dejong, 2013. Developing a Carbon Budget, Physiology, Growth, and Yield Potential Model for Almond Trees. Project Final Report, Plant Sciences Department, UC Davis.
- Turner, D.P., W.D. Ritts, W.B. Cohen, S.T. Gower, S.W. Running, M. Zhao, M.H. Costa, A. Kirschbaum, J. Ham, S. Saleska, and D.E. Ahl. 2006. Evaluation of MODIS NPP and GPP products across multiple biomes. *Remote Sensing of Environment* 102: 282-292.
- van Kooten, G. C., Eagle, A. J., Manley, J., & Smolak, T., 2004. How costly are carbon offsets? A meta-analysis of carbon forest sinks. *Environmental Science & Policy*, 7(4), 239-251.
- Vashum, K. T., & Jayakumar, S., 2012. Methods to estimate above-ground biomass and carbon stock in natural forests-a review. *Journal of Ecosystem & Ecography*, 2(116), 1-7.
- Warner, E., Inman, D., Kunstman, B., Bush, B., Vimmerstedt, L., Peterson, S., Macknick, J., & Zhang, Y., 2013. Modeling biofuel expansion effects on land use change dynamics. *Environmental Research Letters*, 8(1), 015003.

- West, T. O., Brandt, C. C., Baskaran, L. M., Hellwinckel, C. M., Mueller, R., Bernacchi, C. J., Bandaru, V., Yang, B., Wilson, B. S., Marland, G., Nelson, R. G., De La Torre Ugarte, D. G., & Post, W. M. 2010. Cropland carbon fluxes in the United States: increasing geospatial resolution of inventory-based carbon accounting. *Ecological Applications*, 20(4), 1074-1086.
- Wilson, B. T., Woodall, C. W., & Griffith, D. M., 2013. Imputing forest carbon stock estimates from inventory plots to a nationally continuous coverage. *Carbon Balance Management*, 8(1).
- Woodall, C. W., Domke, G. M., Macfarlane, D. W., & Oswalt, C. M., 2012. Comparing field-and model-based standing dead tree carbon stock estimates across forests of the US. *Forestry*, 85(1), 125-133.
- Woodbury, P. B., Smith, J. E., & Heath, L. S., 2007. Carbon sequestration in the US forest sector from 1990 to 2010. *Forest Ecology and Management*, 241(1), 14-27.
- Woodwell, G. M., Hobbie, J. E., Houghton, R. A., Melillo, J. M., Moore, B., Peterson, B. J., & Shaver, G. R., 1983. Global deforestation: contribution to atmospheric carbon dioxide. *Science*, 222(4628), 1081-1086
- Wulder, M. A., White, J. C., Goward, S. N., Masek, J. G., Irons, J. R., Herold, M., Cohen, W. B., Loveland, T. R., & Woodcock, C. E., 2008. Landsat continuity: Issues and opportunities for land cover monitoring. *Remote Sensing of Environment*, 112(3), 955-969.
- Xu, L., & Baldocchi, D. D., 2004. Seasonal variation in carbon dioxide exchange over a Mediterranean annual grassland in California. *Agricultural and Forest Meteorology*, 123(1), 79-96.
- Yang, X., Tang, J., Mustard, J. F., Lee, J. E., Rossini, M., Joiner, J., Munger, J. W., Kornfeld, A., & Richardson, A. D., 2015. Solar-induced chlorophyll fluorescence that correlates with canopy photosynthesis on diurnal and seasonal scales in a temperate deciduous forest. *Geophysical Research Letters*.